

1 **Biosynthesis of SiO₂ nanoparticles using extract of *Nerium oleander***
2 **leaves for the removal of tetracycline antibiotic**
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32 **Abstract**

33 Tetracycline (TC) is one of the antibiotics that is found in wastewaters. TC is toxic,
34 carcinogenic, and teratogenic. In this study, the tetracycline was removed from water by
35 adsorption using dioxide silicon nanoparticles (SiO₂ NPs) biosynthesized from the extract of
36 *Nerium oleander* leaves. These nanoparticles were characterized using SEM-EDX, BET-BJH,
37 FTIR-ATR, TEM, and XRD. The influences of various factors such as pH solution, SiO₂ NPs
38 dose, adsorption process time, initial TC concentration, and ionic strength on adsorption
39 behaviour of TC onto SiO₂ NPs were investigated. TC adsorption on SiO₂ NPs could be well
40 described in the pseudo-second-order kinetic model and followed the Langmuir isotherm
41 model with a maximum adsorption capacity was 552.48 mg/g. At optimal conditions, the
42 experimental adsorption results indicated that the SiO₂ NPs adsorbed 98.62% of TC. The
43 removal of TC using SiO₂ NPs was 99.56% at conditions (SiO₂ NPs dose=0.25 g/L, C₀=25
44 mg/L, and t=40 min) based on Box–Behnken design (BBD) combined with response surface
45 methodology (RSM) modelling. Electrostatic interaction governs the adsorption mechanism is
46 attributed. The reusability of SiO₂ NPs was tested, and the performance adsorption was 85.36%
47 after the five cycles. The synthesized SiO₂ NPs as promising adsorbent has a potential
48 application for antibiotics removal from wastewaters.

49 **Keywords:** Green synthesis; Tetracycline; SiO₂ nanoparticles; *Nerium oleander* leaves extract;
50 Chemical engineering; Adsorption and Optimization

51

52 **1 Introduction**

53 The excessive accumulation of toxic antibiotics in natural aquatic systems has received
54 enormous attention due to their acute toxicity and possible carcinogenic effect (Grenni et al.,
55 2018). Various antibiotics such as quinolones, sulfonamides, macrolides, and tetracyclines
56 are broadly used to prevent and treat infectious diseases (Ahamad et al., 2019). Antibiotic

57 residues, frequently found in soil, sediment, and aquatic environments, have adverse side
58 effects as bacterial resistance changes in the microbial ecological functions (Sodhi et al., 2021).
59 Antibiotic accumulation may also severely impair human physiological functions and have
60 carcinogenic, teratogenic, or hormonal effects (Liu et al., 2017); their excess entrance into food
61 chains may cause different disorders in the human body, such as the gastrointestinal system
62 (Rashidi Nodeh et al., 2020). Therefore, controlling and handling antibiotic contaminants is
63 necessary to have a safe environment (Yu et al., 2016).

64

65 Tetracycline (TC) is considered one of the most dangerous antibiotics due to its toxic,
66 carcinogenic, and mutagenic effects (Scaria et al., 2021). TC is commonly used for animal
67 husbandry and poultry industries worldwide to promote animal growth and prevent infections
68 (Song et al., 2020; Van et al., 2020). About 60-80 % of tetracyclines were subject to different
69 natural surroundings original in or metabolized forms due to lower metabolic rates in animals
70 and humans (Epps and Blaney, 2016; Scaria et al., 2021); due to the high solubility rate, TC is
71 detected in different water bodies and some regions; its concentration exceeds standard
72 environmental limits (Daghrir and Drogui, 2013). Several approaches such as photocatalysis
73 (Zhu et al., 2013), flocculation (Fu et al., 2015), biological treatment (Belkheiri et al., 2011),
74 electrochemical (Wang et al., 2021), adsorption (Sharma et al., 2020), and reverse osmosis
75 (Rostam and Taghizadeh, 2020) have been used to remove tetracycline from wastewaters. The
76 adsorption process is a competitive and practical method for removing TC from wastewater
77 due to high efficiency, low energy demand, and the possibility of reusing adsorbent materials
78 (El Khomri et al., 2020; El Messaoudi et al., 2016a).

79

80 Consequently, a wide range of adsorbent materials (clays, chitosan, coated silica gel, zeolite
81 ionic liquids, metal oxides, nanocomposites, nanomaterials, industrial and agricultural wastes)

82 have already been tested for the recovery of TC contained in wastewater (Chang et al., 2012;
83 Gao et al., 2012; Hao et al., 2021; Maged et al., 2020; Zhou et al., 2017). Recently, conducting
84 nanoparticles such as ZrO₂ (Debnath et al., 2020), NiFe (Ravikumar et al., 2019), CeO₂
85 (Nurhasanah et al., 2020), La₂S₃ (Rashidi Nodeh et al., 2020), and ZnO (Bembibre et al., 2022)
86 have been categorized as efficient adsorbents for TC removal from wastewaters due to high
87 stability, porous nature, good adsorption aptitude, simple doping/dedoping, and ion exchange
88 properties. Many methods have been used to synthesize the SiO₂ nanoparticles in the literature,
89 such as solid-state, sonochemical (Masjedi-Arani et al., 2015), hydrothermal (Potapov et al.,
90 2020), sol-gel (Dubey et al., 2015), and ultrasonic-assisted approaches (Ullah et al., 2019).

91

92 In this study, we used SiO₂ nanoparticles to remove TC from the aqueous solution because
93 they are an excellent adsorbent to remove dyes and metals and a photocatalyst to treat
94 wastewaters (Hosseini et al., 2018; Sharma et al., 2021). These nanoparticles are
95 biosynthesized from the extract of *Nerium oleander* leaves. The biosynthesis of nanoparticles
96 from the extract of agricultural solid wastes is regarded as a simple, rapid, cost-effective, and
97 eco-friendly method for creating nanostructured materials such as metals, metal chalcogenides,
98 and bimetal oxide (Das et al., 2018). In this work, we used the extract of *Nerium oleander*
99 leaves for biosynthesized of SiO₂ nanoparticles. We chose this agricultural solid waste
100 because its extract contains a significant amount of minerals, lost cost materials, and
101 is eco-friendly and abundant in Morocco (Martín et al., 2018; Sebeia et al., 2019).

102

103 The performance of SiO₂ NPs was evaluated in this study by removing tetracycline from the
104 solution aqueous. Besides, characterization techniques were practiced for investigating its
105 structure and properties, such as scanning electron microscope coupled with energy-dispersive
106 X-ray (SEM-EDX), Brunauer-Emmett-Teller and Barrett-Joyner-Halenda (BET-BJH), Fourier

107 transform infrared spectroscopy coupled with attenuated total reflectance (FTIR-ATR),
108 transition electron microscope (TEM), X-ray diffraction (XRD), and point of zero charge
109 (PZC). The present work will substantially impact as it will add unique knowledge on the
110 adsorption potential of SiO₂ NPs synthesized from the extract of *Nerium oleander* leaves. In
111 this current study, the SiO₂ NPs were synthesized from the *Nerium oleander* leaves extract.
112 Different influencing experimental parameters such as pH, SiO₂ NPs dose, reaction time, TC
113 concentration, and ionic strength were investigated. Kinetic and equilibrium models evaluated
114 the adsorption of TC on SiO₂ NPs. The removal of TC optimized using Box–Behnken design
115 (BBD) combined with response surface methodology (RSM). The adsorption mechanism of
116 TC molecules on the SiO₂ NPs surface was proposed. The reusability of SiO₂ NPs was
117 evaluated.

118 **2 Experimental**

119 **2.1 Materials**

120 *Nerium oleander* leaves were collected in Tinghir (South-East Morocco). The silicate of
121 sodium (Na₂SiO₃), tetracycline (antibiotic, C₂₂H₂₄N₂O₈, MW=444.435 g/mol), C₂H₅OH, HCl,
122 and NaOH were purchased from Sigma-Aldrich. The distilled and deionized waters were used
123 through experiments.

124 **2.2 Biosynthesis of SiO₂ NPs**

125 10 g of *Nerium oleander* leave powder were added in 100 mL of C₂H₅OH and stirred for 3 h.
126 After that, the solution was filtered using filter paper then the above mixture was centrifuged
127 for obtained a clear extract of *Nerium oleander* leaves. 2 g Na₂SiO₃ and 50 mL of extract and
128 5 mL of NaOH (0.5 M) were mixed and stirred for 4 h (Dobrucka and Długaszewska, 2016).
129 After precipitation, the mixture was filtered and centrifuged for separated the liquid and the
130 precipitate (SiO₂ NPs). The residue obtained was washed with deionized water. Finally, The

131 SiO₂ nanoparticles were over-dried (80 °C) for 24 h and calcined in the furnace at 500 °C for
132 3 h (Shaligram et al., 2009).

133 **2.3 Characterization**

134 The physicochemical properties of SiO₂ NPs were examined using different characterization
135 techniques. The morphology and microstructures of SiO₂ NPs were examined by SEM-EDX
136 analysis (JEOL, JSM-IT200) and TEM analysis (Philips CM-30). BET and BJH methods
137 (Belsorp Mini II) were used to determine the surface area, total pore volume, and diameter
138 pore of SiO₂ NPs. The chemical bond characteristics of SiO₂ NPs before and after TC
139 adsorption was acquired by FTIR-ATR analysis (Jasco 4100). The SiO₂ NPs crystal structures
140 were evaluated using XRD analysis recorded on a 6100-Shimadzu. The PZC of adsorbent was
141 determined using the method reported by Fiol and Villaescusa, (2009).

142 **2.4 Batch adsorption experiments**

143 The TC adsorption onto SiO₂ NPs nanoparticles was conducted in batch mode using 12.5 mg
144 of SiO₂ NPs in 50 mL of TC solutions with concentrations varied from 50 to 200 mg/L at 23±1
145 °C. The influence of pH solution on adsorption was assessed and ranged from 3 to 11 and was
146 adjusted by 0.01 M HCl acid or 0.01 M NaOH. The adsorbent dose, kinetic reaction, and ionic
147 strength were changed from 0.05 to 0.4 g/L from 5 to 120 min and from 0 to 0.4 M,
148 respectively. The separation of the solid-liquid phases was performed by centrifuging. The
149 residual concentration was determined using a UV/Vis spectrophotometer (2300/Techcomp)
150 at 376 nm as λ_{\max} of TC. The quantity adsorbed q_e (mg/g) and the and TC removal efficiency
151 (%) were obtained using following Eqs. (1) and (2) respectively (El Messaoudi et al., 2016b):

$$152 \quad q_e = \frac{(C_0 - C_e) \times V}{w} \quad (1)$$

$$153 \quad \text{Percentage Removal (\%)} = \frac{(C_0 - C_e)}{C_0} \times 100 \quad (2)$$

154 In equations (1) and (2), C_0 (mg/L) and C_e (mg/L) denote the concentrations of TC before and
155 after adsorption, respectively. The w (g) and V (L) represents the weight of SiO₂ NPs and the
156 volume of the reaction solution, respectively (Bentahar et al., 2018; El Messaoudi et al., 2017).

157 **2.5 Experimental design**

158 BBD used a static method to design experimental parameters influencing the adsorption of TC
159 on SiO₂ NPs using design-expert software (version 12.0.3). TC concentration (A), reaction
160 time (B), and SiO₂ NPs dose (C) as factors have significant effects on the TC adsorption on
161 SiO₂ NPs at 23±1 °C and pH=5. Design of experiment runs and corresponding responses for
162 TC removal efficiency by SiO₂ NPs are summarized in **Table 1**. A three-factors and levels (–
163 1, 0, and 1) were applied to 21 experiments. The TC removal efficiency R (%) was expressed
164 using the quadratic polynomial model was formalized in Eq. (3) (Jawad et al., 2020):

$$165 \quad R(\%) = \sum_{I=1}^3 \delta_i X_i + \sum_{I=1}^3 \delta_{ii} X_i^2 + \sum_{I<J} \delta_{ij} X_i X_j + \delta_0 \quad (3)$$

166 where δ_0 denotes constant-coefficient, δ_i is attributed to the direct effect, δ_{ii} corresponds to
167 higher-order effect, and δ_{ij} denote reciprocal effect.

168 **3 Results and discussion**

169 **3.1 Material characterization**

170 The microstructures of synthesized nanoparticles were analyzed using SEM coupled with
171 EDX. The results obtained are shown in **Fig. 1**. According to **Fig. 1(a)**, The SEM image of
172 SiO₂ NPs showed that the particles are agglomerated and have poor dispersion. **Fig. 1(b)**
173 represents the EDX spectrum and elemental analysis of SiO₂ NPs. The synthesis of SiO₂ NPs
174 confirmed by the presence of O (53.82%) and Si (46.18%) (Dubey et al., 2015). **Fig. 1(c)** and
175 **(d)** indicate the uniform distribution of O and Si was illustrated by EDX elemental mapping.
176 **Fig. 2** illustrates the N₂ isotherms adsorption-desorption and average diameter distribution for
177 SiO₂ NPs nanospheres. By means, BET and BJH methods, the obtained middle surface area,

178 pore diameter, and total pore volume were 583.46 m²/g, 3.46 nm, and 0.27 cm³/g, respectively,
179 confirmed the porosity of SiO₂ NPs. The FTIR spectra of SiO₂ NPs before and after the
180 adsorption of TC (TC-SiO₂) are depicted in **Fig. 3**. On the spectrum of SiO₂ NPs, the broad
181 bands at 3452 cm⁻¹ and 1634 cm⁻¹ correspond to OH stretching vibration and absorbed
182 water molecule, respectively (El Messaoudi et al., 2016a; Niksefat et al., 2014).

183
184 The peaks at 1083 cm⁻¹, 953 cm⁻¹ and 806 cm⁻¹ and 457 cm⁻¹ were ascribed to the Si-O
185 stretching vibration, Si-OH stretching vibration and Si-O-Si symmetric stretching (Rafigh and
186 Heydarinasab, 2017), and Si-O-Si bending, respectively (Yue et al., 2019), which confirms
187 successful synthesis of SiO₂ nanospheres. The spectrum of TC-SiO₂ NPs shows small changes
188 that demonstrate the TC adsorption on the surface of SiO₂ NPs. **Fig. 4(a)** shows the
189 nanoparticles of SiO₂ NPs using a TEM image. According to this figure, the synthesized SiO₂
190 NPs particles were found to be spherical in structure (Dubey et al., 2015; Nita et al., 2019). The
191 XRD characterization results of SiO₂ NPs are provided in **Fig. 4(b)**. A broad peak in the range
192 of 20–30° is corresponded to the crystallite of SiO₂ particles (Rafigh and Heydarinasab, 2017),
193 indicating the successful synthesis of SiO₂ NPs.

194 **3.2 Adsorption study**

195 **3.2.1 Effect of solution pH**

196 **Fig. 5(a)** presents the influence of pH on the TC adsorption at pH values ranging from 3 to 11.
197 This experiment attempted 10 mg of SiO₂ NPs in 50 mL of the solution TC (50 mg/L) at T=23
198 ±1 °C for 120 min. As **Fig. 5(a)** shows, the highest removal of TC was 98.07% at pH=5.
199 Similar results were obtained by Debnath et al. (2020) and Rashidi Nodeh et al. (2020). This
200 increase of TC removal in the acidic medium can be explained by the charge positive of SiO₂
201 NPs and the charge negative of TC (TCH⁻, TC⁻) (Li et al., 2010; Mohammed and Kareem,
202 2019). As **Fig. 10(a)** depicted, PZC of SiO₂ NPs was 8.2. The quantity adsorbed was 196.17

203 mg/g at pH=5. The dominance of charge positive of SiO₂ NPs at pH<PZC (adsorption of TC
204 was favourable) and negatively charged when pH>PZC (adsorption of TC was unfavourable)
205 (Song et al., 2020).

206 **3.2.2 Effect of SiO₂ NPs dose**

207 A TC concentration of 50 mg/L with a pH=5 at T=23 ±1 °C for 120 min, the effect of SiO₂
208 NPs dose (0.05-0.4 g/L) on TC adsorption was studied. **Fig. 5(b)** shows the results obtained.
209 The TC removal increased from 54.38 to 98.92% by increasing the SiO₂ NPs dose from 0.05
210 to 0.25 g/L, while the quantity adsorbed decreased from 543.88 to 197.95 mg/g. Results
211 imply that the number of active adsorption sites for TC adsorption corresponds to the applied
212 dose, prompting higher removal efficiency (Zhang et al., 2019). After equilibrium between
213 the adsorbent and antibiotic solution, the removal percentage remains consistent at higher
214 dosages (> 0.25 g/L) (Jin et al., 2019). The optimum adsorbent dosage was considered 0.25
215 g/L to reach maximum TC removal efficiency, respectively.

216 **3.2.3 Effect of contact time**

217 The influence of the contact time on TC adsorption using SiO₂ NPs displayed in **Fig. 5(a)**. The
218 contact time ranged from 5 to 120 min, whereas other parameters were kept constant (SiO₂
219 NPs=0.25 g/L, TC concentration=50 mg/L, pH=5, and T=23±1 °C). The TC adsorption was
220 fast at first 30 min, which may be attributed to many sites accessible on the surface of the
221 SiO₂ NPs in the initial phase (Ahamad et al., 2019). The equilibrium time was found to be
222 40 min. After equilibrium adsorption, the active sites were occupied by the TC
223 molecule/ions. Therefore the adsorption rate became consistent (Zhou et al., 2020).
224 Experimental data showed that stability was achieved in 40 min with an adsorption capacity
225 of TC was 195.97 mg/g.

226 **3.2.4 Effect of Initial TC concentration**

227 The influence of TC concentration (25–200 mg/L) on its retention using SiO₂ NPs was studied
228 with a fixed pH=5 and 0.25 g/L of SiO₂ NPs dose for 40 min. As illustrated in **Fig. 5(d)**, by
229 increasing TC concentration from 25 to 150 mg/L, the TC adsorption capacity progressively
230 increased from 99.40 to 512.06 mg/g, while the removal of TC decreased from 99.40 to
231 85.34%. This increases TC adsorption capacity due to the occupation of all available sites on
232 the surface of SiO₂ nanoparticles by TC molecules (Debnath et al., 2020). After 150 mg/L, a
233 plateau was not achieved in the adsorption capacity, suggesting active sites are still available
234 and no saturation occurred. The phenomena were similar to the adsorption research of
235 tetracycline reported previously (Ahamad et al., 2019; Ravikumar et al., 2019).

236 **3.2.5 Effect of ionic strength**

237 The electrolyte (NaCl) concentration in the aqueous solution significantly affects TC
238 adsorption onto SiO₂ NPs. As reported in **Fig. 6**, at SiO₂ NPs=0.25 g/L, TC
239 concentration=50 mg/L, pH=7, and T=23±1 °C, the removal efficiency of SiO₂ NPs for TC
240 decreased from 73.15% (0 M NaCl) to 55.48% (0.4 M NaCl) by increasing the electrolyte
241 concentrations from 0 to 0.4 M NaCl. By increasing the ionic strength, Cl⁻ competes
242 with negatively charged TC (TCH⁻, TC⁻) species for adsorption onto SiO₂ NPs surface site
243 with positively charged at pH=7 (Zhang et al., 2019). This result is in line with another study
244 that reported a decline in TC adsorption onto clay surface sites, negatively charged by
245 increasing electrolyte concentration in the solution (Parolo et al., 2008). In their study, by
246 increasing Na⁺ ion concentration at pH=4 of the solution, the competence for occupying surface
247 sites between Na⁺ ion and positively charged TC (TCH⁺) increased (Yang et al., 2011). The
248 reduction in TC adsorption onto the surface of clay increased in a higher concentration of Na⁺
249 (Parolo et al., 2008).

250 **3.3 Adsorption kinetics**

251 The obtained experimental data were evaluated by using the pseudo-first-order (PFO) (Simonin,
252 2016), pseudo-second-order (PSO) (Ho and McKay, 1998), and intraparticle diffusion (IPD)
253 (Graaf et al., 1990) kinetic models and their linear forms are given using the following Eqs. (4),
254 (5), and (6) as represented in **Table 2**, respectively. **Table 2** also shows the parameters for linear
255 fitting. Furthermore, the PSO model was fitted to data experimental based on R^2 (correlation
256 coefficient). It can be found that R^2 values are very near to 1, and $q_{e,exp}$ values are also closer to
257 $q_{e,cal}$ values for the PSO. This model speculates that adsorption pursues a second-order mechanism
258 (El Messaoudi et al., 2021b).

259 **3.4 Adsorption isotherm**

260 The TC adsorption on SiO₂ NPs was the Langmuir (Langmuir, 1918), Freundlich (Freundlich,
261 1907), and Temkin (Johnson and Arnold, 1995) isotherm models. The linear forms of these
262 isotherms are expressed on Eqs. (7), (8), and (9), as represented in **Table 3**. The parameters of
263 linear fitting are listed in **Table 3**. Based on regression coefficients R^2 (0.9931, 0.9650, and
264 0.9392), the Langmuir model best described the TC adsorption on SiO₂ NPs. Langmuir model
265 suggests the dye adsorption occurs as a monolayer from the homogeneous surface of the
266 adsorbent (Bentahar et al., 2017; El Messaoudi et al., 2021a). A similar observation was found
267 in other studies, showing successful adsorption isotherm data using Langmuir isotherm
268 compared to Freundlich and Temkin isotherm models (Li et al., 2021; Mohammed et al., 2020).
269 The maximum adsorption capacity Q_m of SiO₂ NPs for TC adsorption was 552.48 mg/g. The
270 comparison of the adsorption of SiO₂ NPs to remove TC from aqueous solution with other
271 adsorbents reported in the literature is summarized in **Table 4**. Based on the results presented in
272 this table, that the SiO₂ NPs exhibit high adsorption of TC compared with adsorbents. Therefore,
273 SiO₂ NPs are a suitable adsorbent for the removal of antibiotic molecules from wastewater
274 treatment.

275 3.5 Design optimization

276 **Table 5** presents the results of ANOVA analysis of the statistical significance. The F-value
277 (162.55) and p-value (<0.0001) indicate the polynomial equation was significant for the
278 removal of TC on SiO₂ NPs within 95% (Hu et al., 2021). The high values of R² (0.9925),
279 adjusted R² (0.9864), and predicted R² (0.9449) indicated the excellent fit of this model to the
280 factors selected (Aziz et al., 2021). The value of adequate precision of 37.9356 (>4) shows a
281 high level of statistical significance (Dalia Allouss et al., 2019). The predicated TC removal
282 using SiO₂ NPs was obtained by the developed model represented below in Eq. (10)
283 (Dalia Allouss et al., 2019).

$$284 R (\%) = 98.57 - 6.84 A + 3.78 B + 0.8861 C + 3.72 AB + 0.6362 AC - 0.0713 BC - 3.57A^2 - 2.36$$
$$285 B^2 - 0.6527 C^2 \quad (10)$$

286 The residual vs. predicted and 3D response surface plots of TC percentage are illustrated in
287 **Fig. 7** and **Fig. 8**, respectively. The TC removal was experimentally 99.56% under optimal
288 conditions (SiO₂ NPs dose=0.25 g/L, C₀=25 mg/L, and pH=5 at 23±1 °C for 40 min) RSM-
289 BBD modeling.

290 3.6 Reusability of SiO₂ NPs

291 To assess the applicability of the adsorbent in the full-scale operation, the reusable capacity of
292 SiO₂ NPs was studied for the removal of TC in optimal conditions (SiO₂ NPs dose=0.25 g/L,
293 t=40 min, C₀=50 mg/L, pH (TC)=5, and T=23±1 °C). Therefore, the regeneration studies for
294 evaluating the adsorption efficiency of the nanocomposite were conducted within five
295 successive cycles using 0.1 M NaOH. According to **Fig. 9**, less than a 13% drop in the removal
296 efficiency of TC occurred at the end of the fifth run, signifying the desirable reusability
297 potential of the synthesized nanoparticles within successive runs of operation (Yang et al.,
298 2020). This decrease was attributed to the occupation of available sites on the SiO₂ NPs surface
299 (El Messaoudi et al., 2021b). In conclusion, the present adsorbent can be regarded as a

300 promising material for practical application in environmental protection due to its excellent
301 adsorption activities and high stability.

302 **3.7 Proposed adsorption mechanism**

303 Various mechanisms are involved in the adsorption of the organic compound to nanoparticle
304 adsorbents. It was reported that both electrostatic and dispersive interaction between adsorbent
305 and adsorbate is important in the adsorption process (Gao et al., 2019). Regarding the properties
306 of organic adsorbate and adsorbent, the importance of each interaction is determined. As **Fig.**
307 **10(a)** shows, PZC of SiO₂ NPs was 8.2. The dominance of charge positive of SiO₂ NPs at
308 pH<PZC and negatively charged when pH>PZC. The high value of PZC indicates the
309 favourable and maximum adsorption in an acidic medium. This result confirmed the effect of
310 pH on adsorption. The adsorption mechanism of TC on SiO₂ NPs is schematized in **Fig. 10(b)**
311 based on PZC. The charge positive of SiO₂ NPs and the negative charge of TC indicates the
312 governance of electrostatic interactions between TC and SiO₂ NPs (Gao et al., 2012).

313 **4 Conclusions**

314 The SiO₂ nanoparticles biosynthesized from the extract *Nerium oleander* leaves successfully
315 with an effective method to remove tetracycline (TC) from an aqueous solution. The result of
316 SEM-EDX, FTIR, TEM, and XRD characterization confirmed the biosynthesis of SiO₂
317 nanoparticles with spherical and crystallite in their structure. The parameters of the adsorption
318 process were optimized with the variation in values of pH solution, SiO₂ NPs dose, adsorption
319 process time, initial TC concentration, and ionic strength. Under conditions (SiO₂ NPs
320 dose=0.25 g/L, $t=40$ min, $C_0=50$ mg/L, pH=5, and $T=23\pm 1$ °C), the TC removal was 98.62%.
321 The kinetics and isotherm of TC adsorption on SiO₂ NPs were described as the PSO and
322 Langmuir models, respectively. The Q_m was 552.48 mg/g. Optimization is an effective
323 approach for modelling the sorption process of the TC on SiO₂ NPs using BBD–RSM. The
324 recyclability study demonstrated that the SiO₂ NPs exhibited excellent reusability for TC

325 removal. These results confirm that the SiO₂ NPs nanoparticles are suitable for removing
326 antibiotics from wastewaters.

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332

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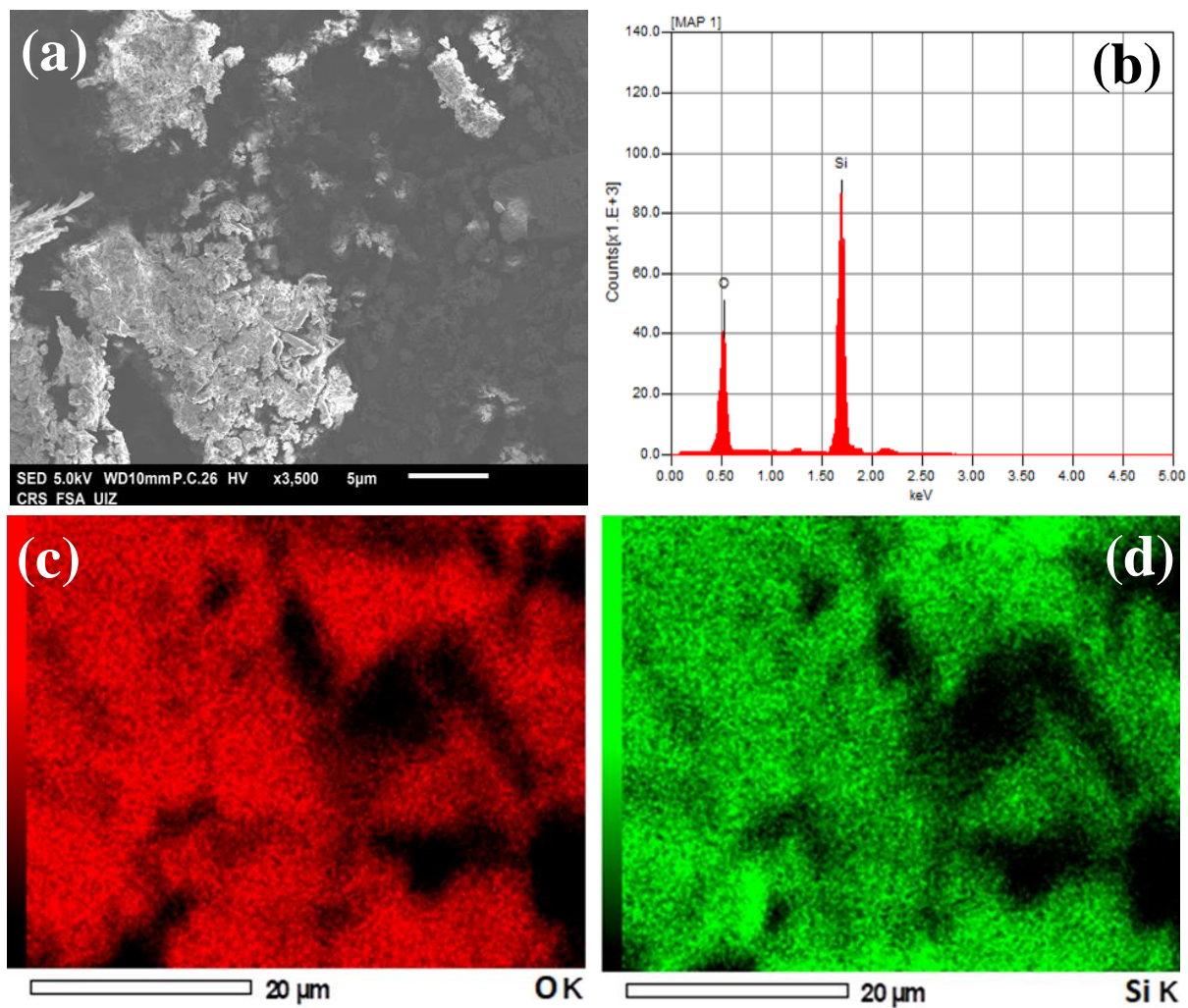
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List of Figures

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Fig. 1. SEM image; (a) SiO₂ NPs, (b) Elemental analysis, (c) Mapping image of SiO₂ NPs: oxygen and (d)

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Mapping image of silicon.

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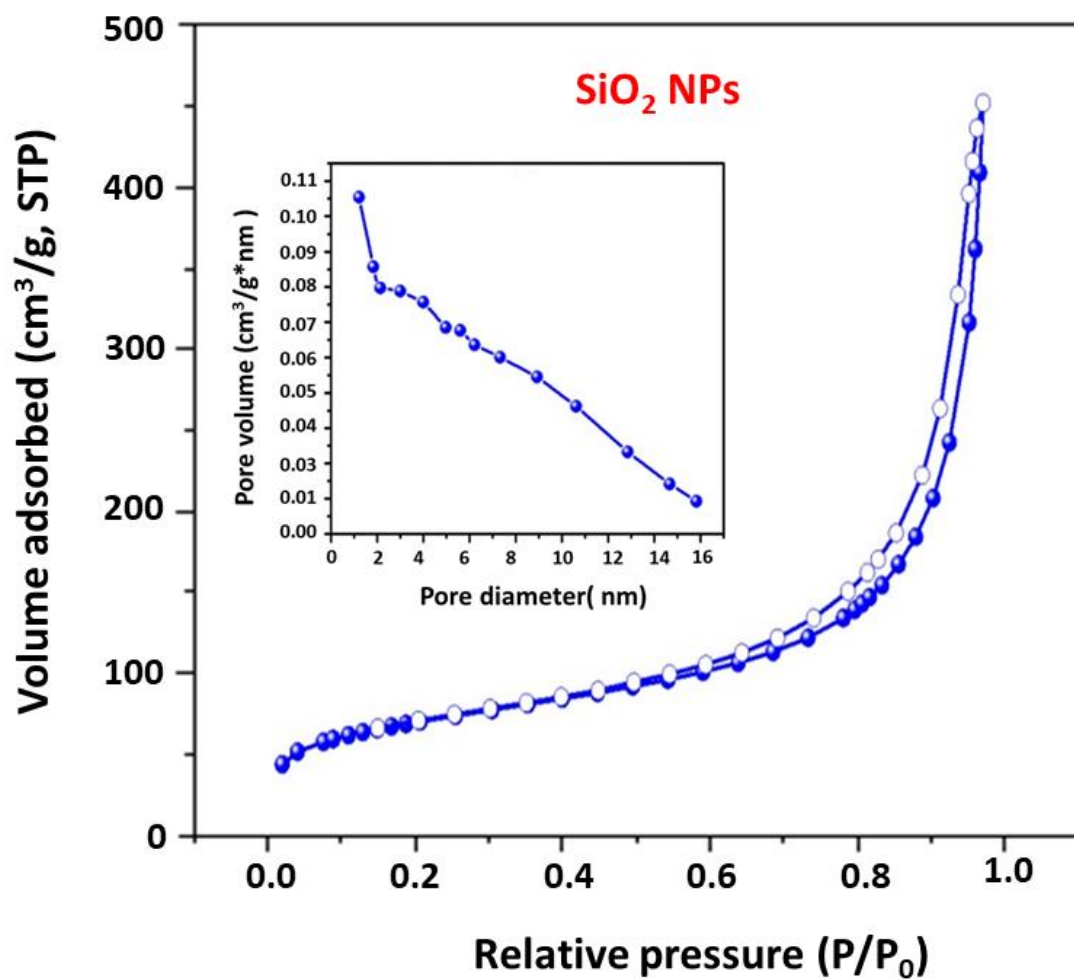


Fig. 2. N₂ adsorption/desorption isotherm curve and pore size distribution of TC-SiO₂ NPs.

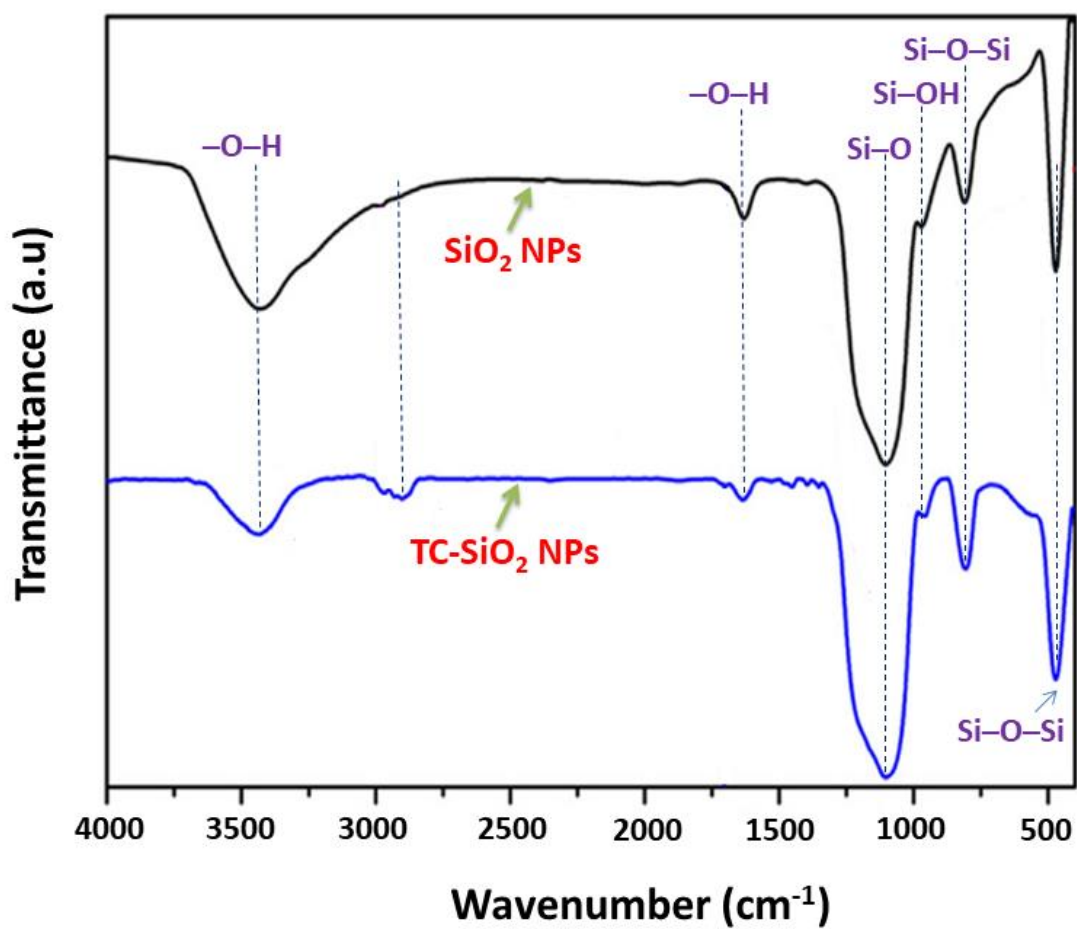


Fig. 3. FTIR spectra of SiO₂ NPs and TC-SiO₂ NPs (after adsorption).

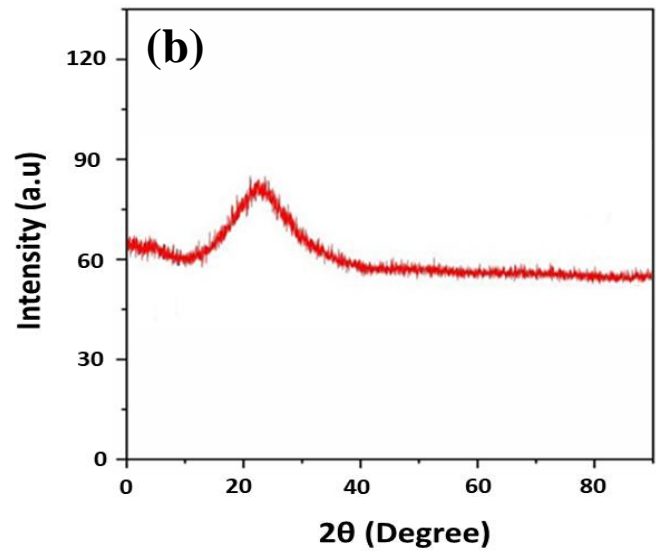
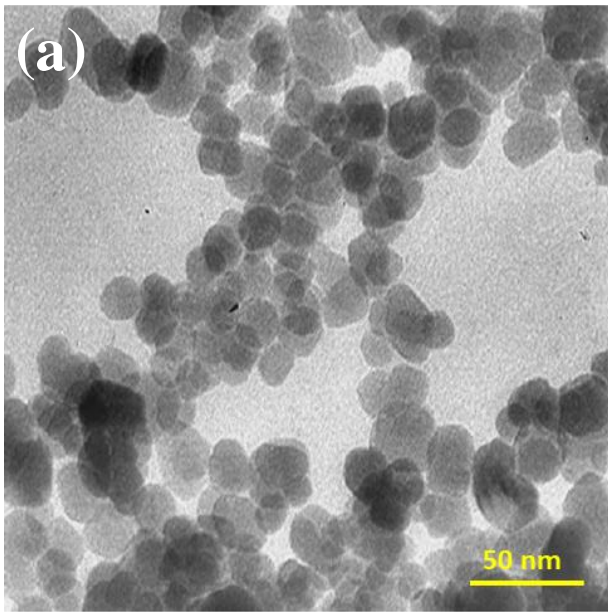
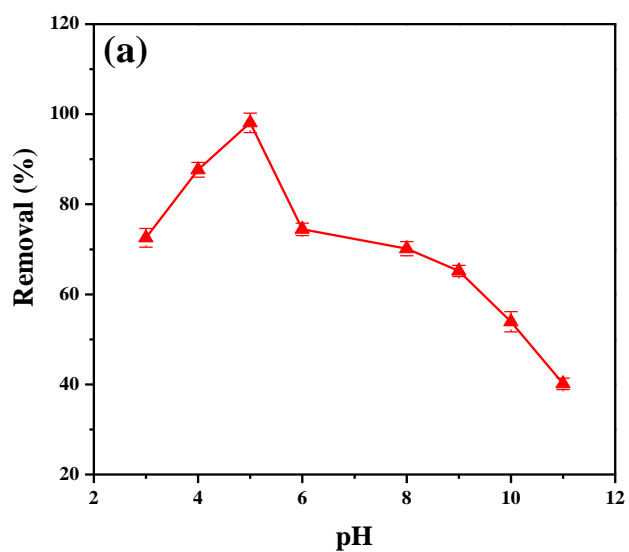
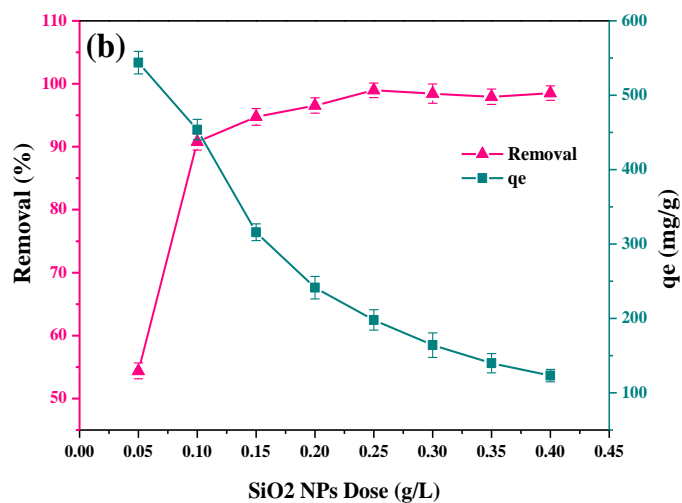


Fig. 4. SiO₂ NPs (a) TEM image and (b) XRD spectrum.

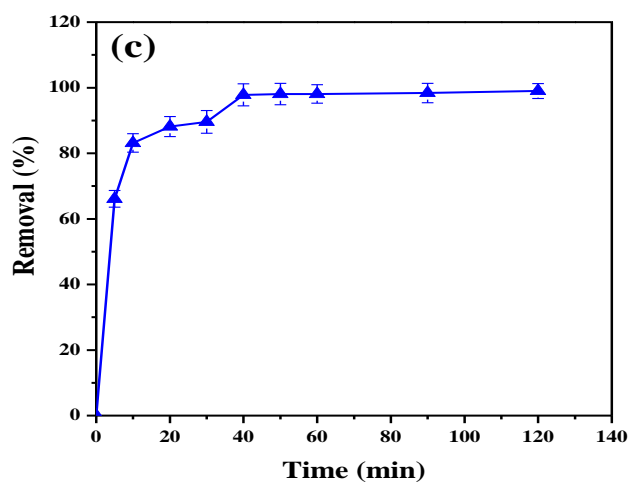
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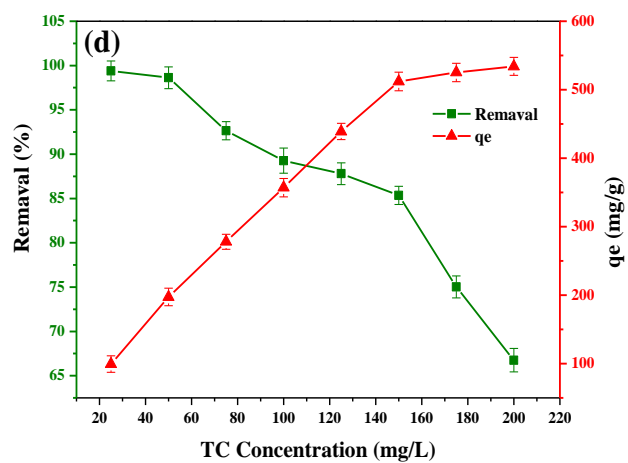
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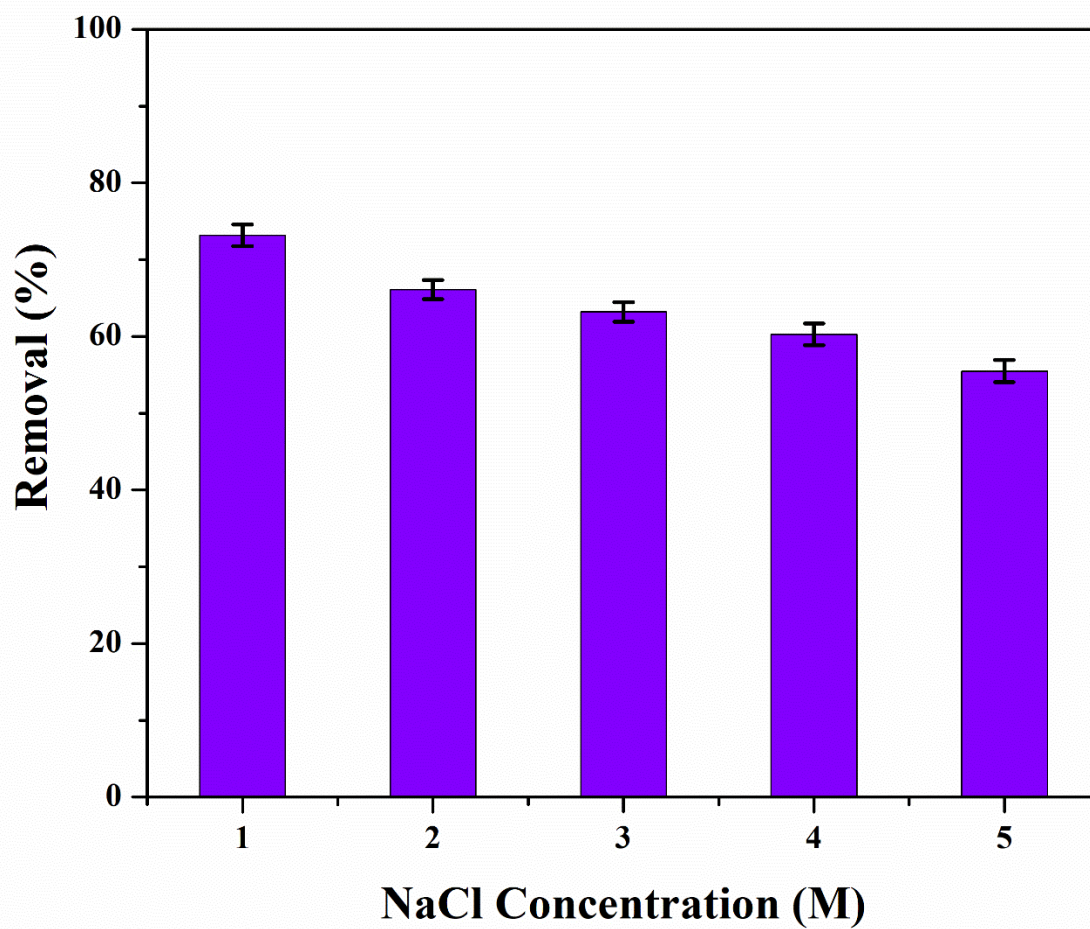
Fig. 5. Removal percentage using SiO₂ NPs; (a) Effects of pH, (b) Effect of SiO₂ NPs dose, (c) Effect of reaction time and (d) Effect of initial concentration on TC adsorption.

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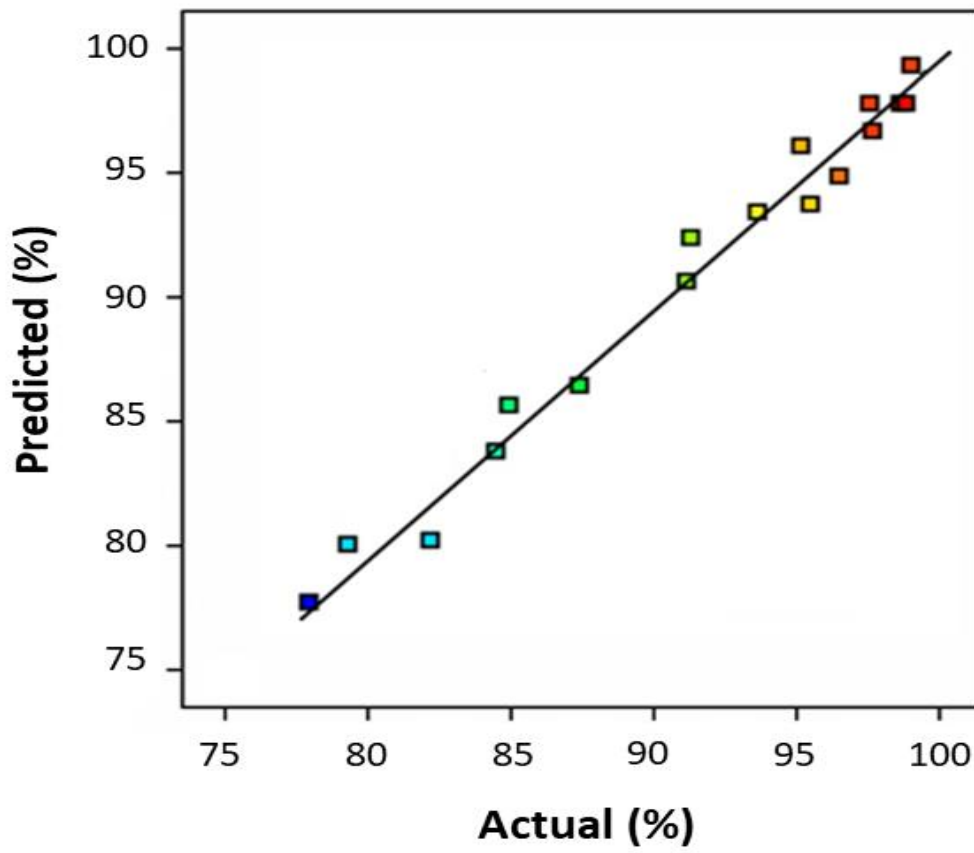
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Fig. 6. Effect of ionic strength on TC adsorption using SiO₂ NPs.

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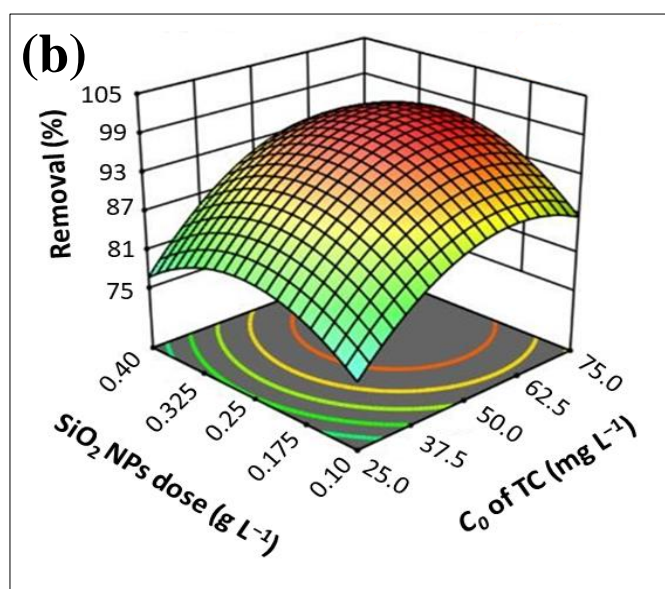
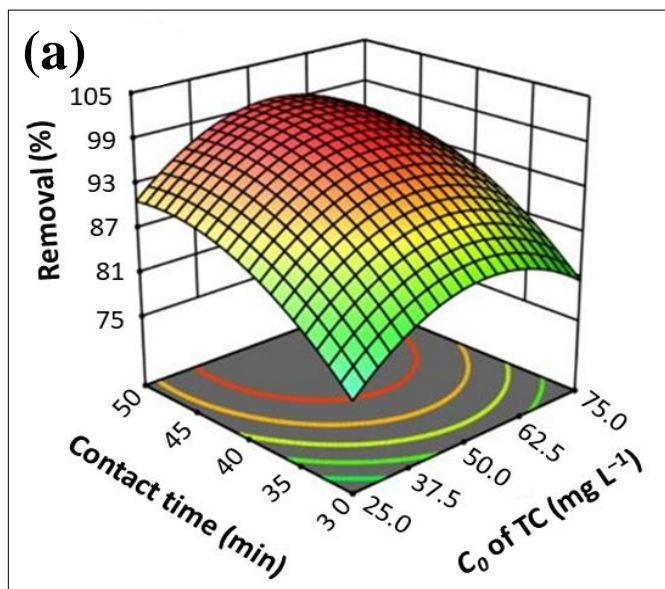
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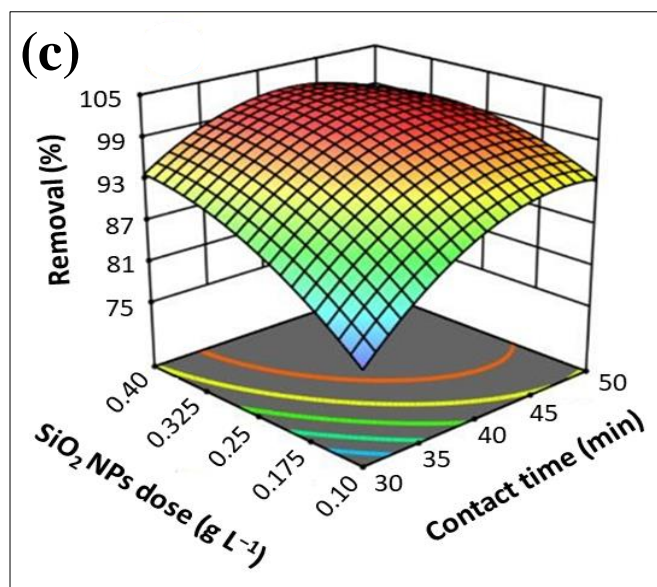
Fig. 7. Actual versus predicted for TC adsorption on SiO₂ NPs.

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Fig. 8. Surface response plots of TC removal percentage; (a) Initial TC concentration with contact time, (b)

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SiO_2 NPs dose with initial TC concentration and (c) SiO_2 NPs dose with contact time.

620

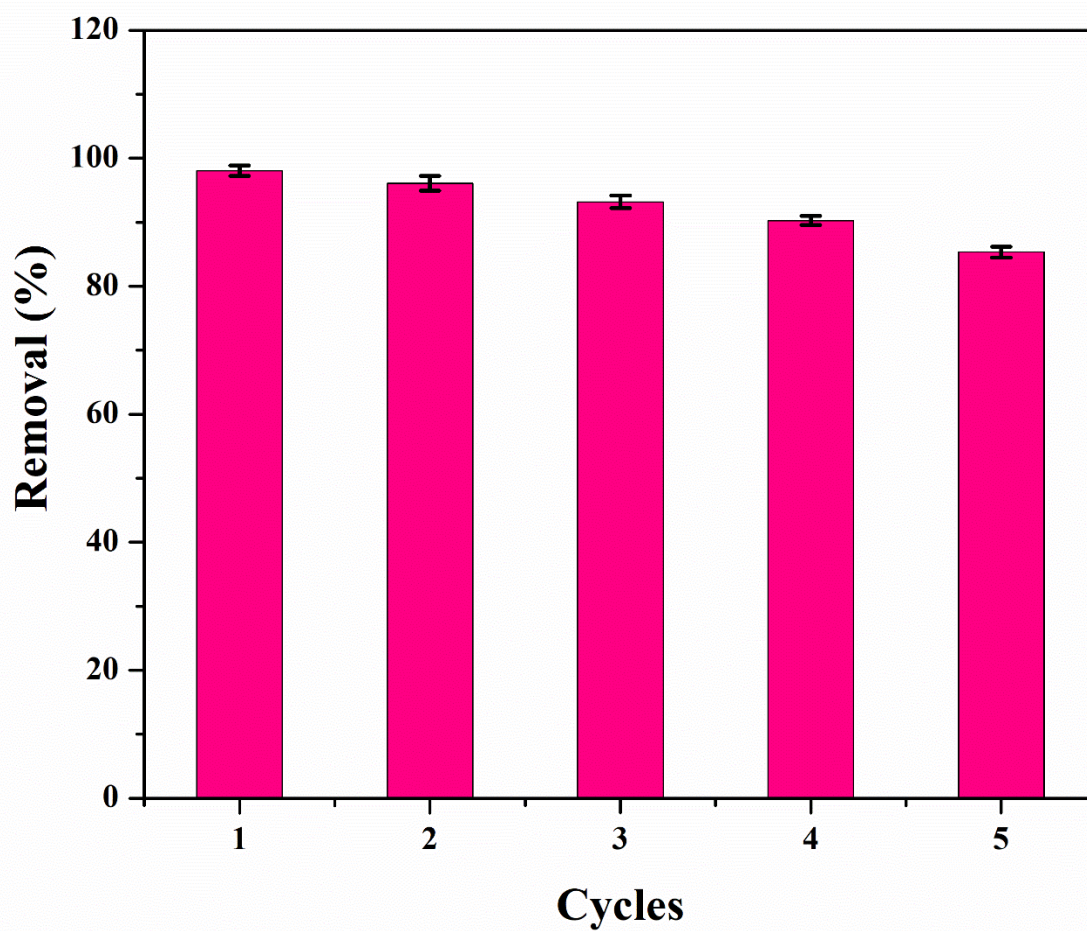
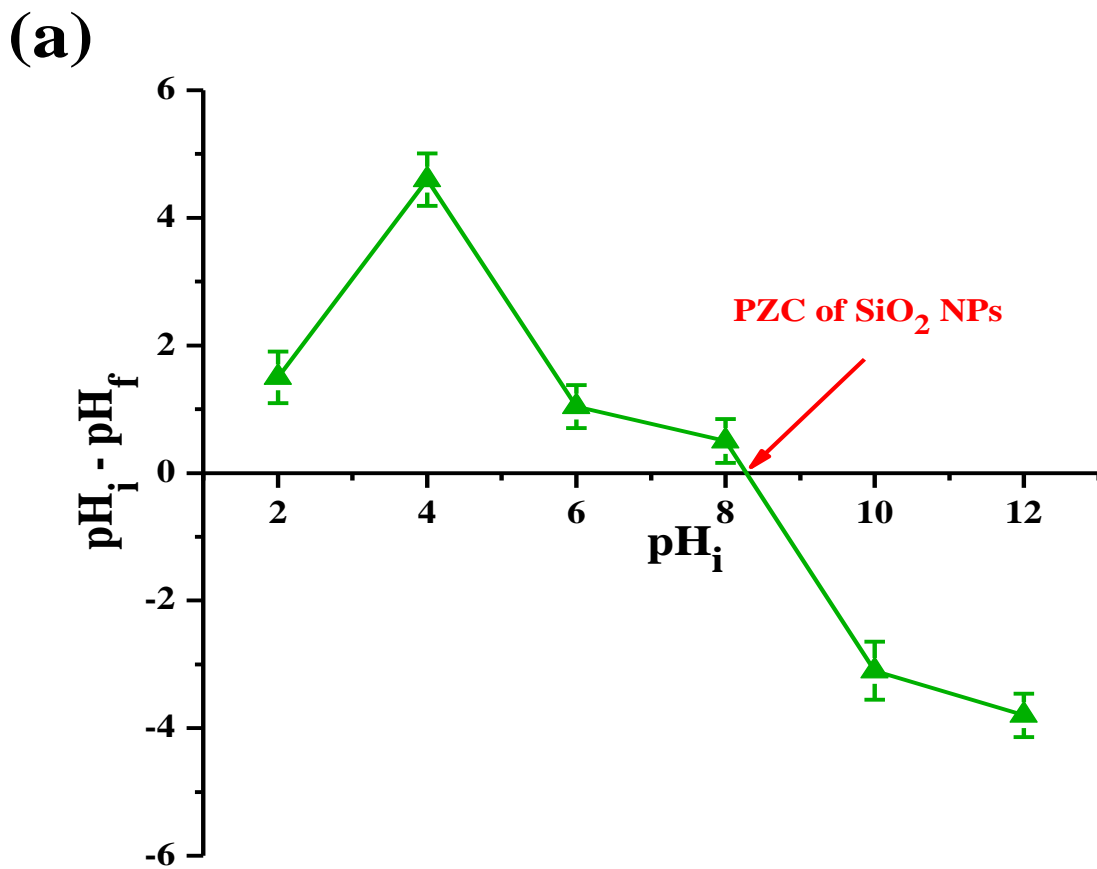


Fig.9. Reusability of SiO₂ NPs for the removal of TC.

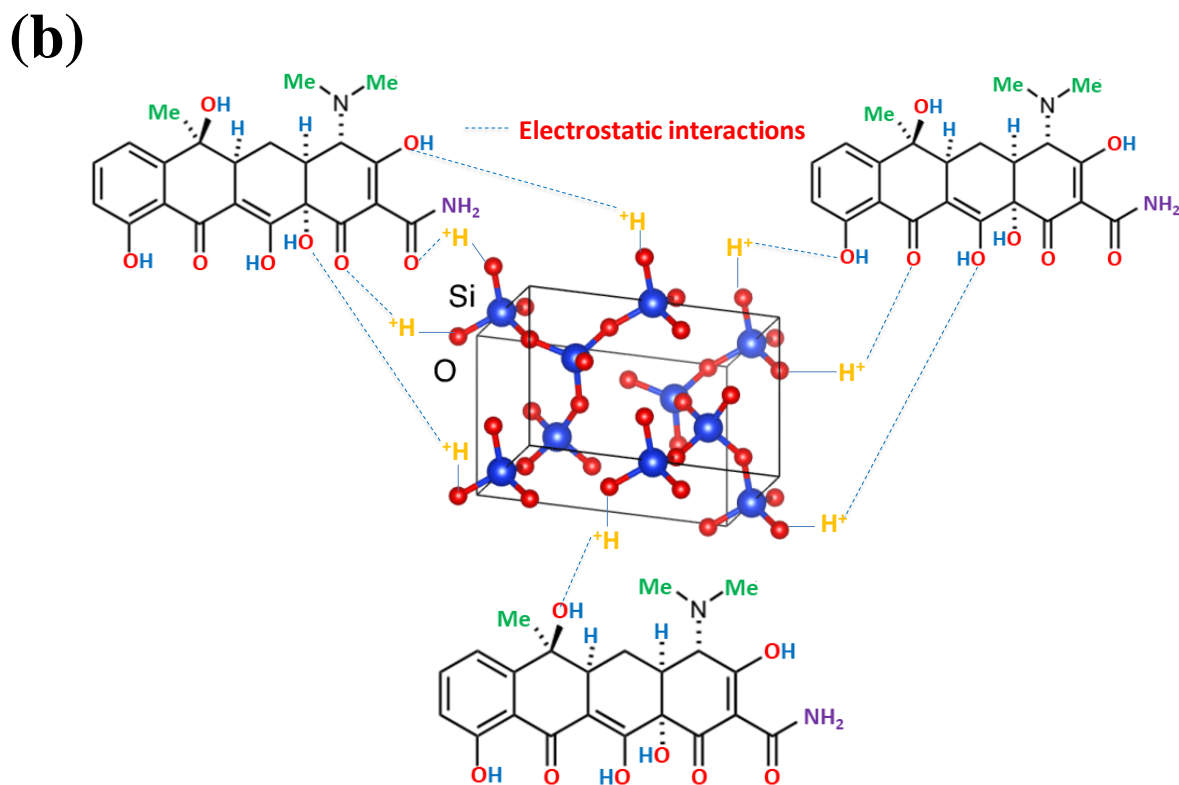
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625

626 **Fig. 10.** (a) PZC of SiO₂ NPs and (b) Proposed adsorption mechanism of TC on SiO₂ NPs.

List of Tables

628 **Table 1.** Design of experiment runs and corresponding responses for TC removal efficiency by
 629 SiO₂ NPs.

Variables		Codes		
		-1	0	1
A=TC concentration (mg/L)		25	50	75
B=Contact time (min)		30	40	50
C=SiO ₂ NPs dose (g/L)		0.1	0.25	0.4
Run	A (mg/L)	B (min)	C (g/L)	R (%)
1	25	30	0.1	98.05
2	75	30	0.1	77.11
3	25	50	0.1	98.87
4	75	50	0.1	90.11
5	25	30	0.4	99.23
6	75	30	0.4	78.13
7	25	50	0.4	99.06
8	75	50	0.4	93.55
9	25	40	0.25	99.56
10	75	40	0.25	90.57
11	50	30	0.25	89.45
12	50	50	0.25	98.66
13	50	40	0.1	90.11
14	50	40	0.4	98.65
15	50	40	0.25	98.44
16	50	40	0.25	98.71
17	50	40	0.25	98.24
18	50	40	0.25	98.53
19	50	40	0.25	98.61
20	50	40	0.25	98.72
21	50	40	0.25	98.23

Table 2. Kinetic model parameters for the TC adsorption onto SiO₂ NPs.

Model and its equation	Parameter	Value
	$q_{e,exp}$ (mg/g)	197.98
PFO	$q_{e,cal}$ (mg/g)	51.03
$Log(q_e - q_t) = Log(q_e) - \frac{K_{PFO}}{2.303} t$ (4)	K_{PFO} (1/min)	0.0052
	R^2	0.8285
PSO	$q_{e,cal}$ (mg/g)	204.08
$\frac{t}{q_t} = \frac{1}{K_{PSO} q_e^2} + \frac{1}{q_e} t$ (5)	K_{PSO} (g/mg.min)	0.0019
	R^2	0.9996
	K_{IPD1} (mg/g.min ^{1/2})	16.0461
	C_1 (mg/g)	35.28
1 st linear portion	R^2	0.9811
	K_{IPD2} (mg/g.min ^{1/2})	10.1487
IPD	2 nd linear portion C_2 (mg/g)	49.76
$q_t = K_{IPD} t^{1/2} + C$ (6)	R^2	0.9045
	K_{IPD3} (mg/g.min ^{1/2})	3.1460
	3 rd linear portion C_3 (mg/g)	62.75
	R^2	0.9733

632

633 **Notation:** K_{PFO} (1/min)=PFO rate constant, K_{PSO} (g/mg/min)=PSO rate constant, K_{IPD}
634 (mg/g.min^{1/2})=IPD rate constant, C (mg/g)=Constant for any experiment, q_t (mg/g)=Amount
635 adsorbed of TC at time t, and q_e (mg/g)=Amount adsorbed of TC at equilibrium (Graaf et al.,
636 1990; Ho and McKay, 1998; Simonin, 2016)

637

638

Table 3. Isotherm model parameters for the TC adsorption onto SiO₂ NPs.

Model and its equation	Parameter	Value
Langmuir	Q_m (mg/g)	552.48
$\frac{C_e}{q_e} = \frac{1}{Q_m K_L} + \frac{C_e}{Q_m} \quad (7)$	K_L (L/mg)	0.3175
	R^2	0.9931
	Freundlich	K_F (mg/g)
$\ln q_e = \ln K_F + \frac{\ln C_e}{n} \quad (8)$	N	03.5997
	R^2	0.9650
	Temkin	K_T (L/g)
$q_e = B \ln K_T + B \ln C_e \quad (9)$	B	75.6710
	R^2	0.9392

640

641 **Notation:** C_e (mg/L)=Equilibrium concentration of TC, Q_m (mg/g)=Monolayer (maximum)
642 adsorption capacity, K_L (L/mg)=Langmuir rate constant, K_F (mg/g)=Freundlich rate constant,
643 n =Heterogeneity factor, K_T (L/g)=Temkin rate constant, and B =constant related to the heat
644 adsorption (Freundlich, 1907; Johnson and Arnold, 1995; Langmuir, 1918)

645

646 **Table 4.** Comparison of the adsorption of SiO₂ NPs for TC with other adsorbents
 647 reported in the literature.

Adsorbent	Ad (mg/L)	T (°C)	t (min)	C ₀ (mg/L)	pH	Q _m (mg/g)	Reference
ZrO ₂ NPs	0.2	–	15	25–150	6	526.32	(Debnath et al., 2020)
AgO/MgO/FeO@Si ₃ N ₄	–	30	90	30–100	8	172.41	(Sharma et al., 2020)
ACCS	02.5	–	20	100–700	5	38.30	(Song et al., 2020)
ZVI@ACCS	02.5	–	20	100–700	5	78.30	(Song et al., 2020)
NiFe ₂ O ₄ @CDs	10	50	1440	25–100	8	591.72	(Liu et al., 2017)
Pristine MoS ₂	0.4	35	2400	50–500	6	409.84	(Li et al., 2021)
NiFe NPs	0.3	–	90	20–80	7	61.00	(Ravikumar et al., 2019)
La ₂ S ₃ NPs	1	25	90	10–300	5	56.81	(Rashidi Nodeh et al., 2020)
SiO ₂ NPs	0.25	23	40	25–200	5	552.48	Current study

648

649 **Notation :** Ad=Adsorbent dose, T=Temperature, t=Time, C₀= Initial TC concentration.

650

651 **Table 5.** ANOVA analysis and regression coefficients for the designed statistical model.

Source	Sun Squares	Df	Mean Square	F-value	p-value
Model	1221.20	9	135.69	162.55	< 0.0001
A-TC concentration	639.35	1	639.35	765.93	< 0.0001
B -Contact time	195.12	1	195.12	233.75	< 0.0001
C- SiO ₂ NPs dose	10.72	1	10.72	12.85	0.0043
AB	110.78	1	110.78	132.71	< 0.0001
AC	03.24	1	03.24	03.88	0.0746
BC	0.0406	1	0.0406	0.0487	0.8295
A ²	191.81	1	191.81	229.78	< 0.0001
B ²	83.37	1	83.37	99.88	< 0.0001
C ²	06.39	1	06.39	07.66	0.0183
Residual	09.18	11	0.8347		
Lack of Fit	08.85	5	01.77	31.76	0.0003
Pure Error	0.3343	6	0.0557		
Core total	1230.38	20			
Model statistics	R ²	Adjusted R ²	Predicted R ²	Adequate precision	
	0.9925	0.9864	0.9449	37.9356	

652