

1 **Title page**

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3 Title of article: **Density-dependent modulation of copepod body size and temperature-size**
4 **responses in a shelf sea** Authors:

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37

38 Running head (max 40 characters): **Density-dependent controls of copepod body size**

39 Keywords (between 5-12): zooplankton; temperature-size response; density dependence;
40 competition for food; climate change; Western Channel Observatory; *Acartia clausi*; *Calanus*
41 *helgolandicus*; *Centropages typicus*; *Oithona similis*; *Paracalanus parvus*; *Pseudocalanus*
42 *elongatus*; *Temora longicornis*.

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49 **Statement of significance**

50 Dear Editor,

51 We hereby submit our manuscript entitled “**Density-dependent modulation of copepod**
52 **body size and temperature-size responses in a shelf sea**” for consideration in *Limnology and*
53 *Oceanography*.

54 Body size is a master trait, strongly influencing many biological rates (feeding, growth,
55 metabolism, reproduction, mortality), and ecological interactions. Body size reduction has also
56 been termed “the third universal response to climatic warming”. The prevailing paradigm is that
57 body size responses of ectotherms to warming are governed largely by temperature and size
58 interactions with “bottom up” environmental factors (oxygen or food). Our field study - of 7098
59 adult female copepods across 13 different seasons spanning a 27-year period of warming at a
60 single shelf sea site - challenges this paradigm. We reveal the importance of density-dependent
61 factors

62 (most likely via competition) in modulating body size. Our study, based on the wellcharacterised
63 Plymouth L4 time series site, allows a far more robust and comprehensive interspecies
64 comparison than is possible, for example, by many meta-analyses which collate patchy
65 heterogeneous data on different species, sites and time periods.

66 For all these reasons, we believe that our study is of wide interest to aquatic ecologists
67 who study ectotherm life-history or population responses to environmental change, including
68 climate warming, as well as to those with special interest in impacts on marine zooplankton. This
69 is why we believe Limnology & Oceanography is the best journal for the publication of our
70 paper. We look forward to hearing your response in due course.

71 **Abstract**

72

73 Body size is a fundamental trait in ecology, and body size reduction with increasing temperature
74 has been termed the third universal response to climate warming. Whereas effects of temperature
75 and food on phenotypic plasticity of zooplankton adult body size have been investigated,

76 density-dependent effects have been neglected. We measured seasonal changes in the prosome
77 length of 7098 adult females of seven dominant copepod species in 13 years spanning a 27-year
78 period of warming at the L4 time series off Plymouth, UK. The seasonal Temperature-Size (T-S)
79 response varied greatly among species, from reductions of 2.93% of carbon mass $^{\circ}\text{C}^{-1}$ for
80 *Paracalanus parvus* to 10.15% of carbon mass $^{\circ}\text{C}^{-1}$ for *Temora longicornis*. Evidence for a
81 longterm T-S response was detected in at least two species, supporting the hypothesis that
82 climatic warming leads to smaller adult sizes. April was a crucial month for determining the
83 strength of the T-S response. During this month, body size related negatively to total
84 zooplankton abundance. We suggest that the mechanism for this density-dependence is via
85 competition for food and/or intraguild predation, since spring was also the period when the ratio
86 of food biomass to zooplankton biomass was at its lowest. Our study is among the first *in situ*
87 demonstrations of density-dependent effects on the body size of marine zooplankton and shows
88 the need to consider the effect of top-down as well as bottom-up factors on body size in a
89 warming climate.

90

91 **Introduction**

92 Body size is a ‘master trait’, influencing a multitude of physiological and ecological processes
93 and rates (Kleiber, 1947; Hirst et al., 2014; Andersen et al., 2016). Understanding how body size
94 is influenced by environmental conditions, including warming, is particularly important, given
95 that trophic interactions and vital rates are highly size-dependent in many pelagic ecosystems
96 (Hansen et al., 1994; Cuthbert et al., 2020). The commonly observed plastic phenotypic
97 response, in which body size at maturity of an ectothermic species declines with increasing

98 developmental temperature, is called the Temperature-Size Rule (Atkinson, 1994). Indeed,
99 smaller adult body size following growth in warmer conditions (a negative TemperatureSize (T-
100 S) response) is evident in many multivoltine aquatic invertebrates in laboratory experiments and
101 across thermally varying seasons and latitudes (Lock and McLaren, 1970; Forster et al., 2012;
102 Horne et al., 2017). More broadly, a warming-induced reduction in body size within a
103 community is regarded as one of the three universal responses to climate change, along with
104 responses of phenology (seasonal timing of life cycles) and geographical redistribution (Gardner
105 et al., 2011).

106

107 Body size reductions with warming are not only observed under controlled experimental
108 temperatures with food satiation (Lock and McLaren 1970; Lee et al., 2003; Forster et al., 2012),
109 but have been widely documented in thermally variable seasonal environments, often at
110 temperate latitudes (Horne et al., 2016, 2017) and with latitude (Horne et al., 2015). Such
111 correlations between seasonally-varying temperature and the body size of adults is largely
112 dependent upon periods in which adult recruitment is significant, and therefore the measured
113 environmental conditions relate closely to those experienced during the juvenile growth phase of
114 the recruiting adults. In many copepod species, in temperate and high latitude environments,
115 recruitment of adults is especially marked between spring and late summer, which has been
116 termed the ‘Temperature Dependent Phase’ (TDP) (followed by a ‘Resting Phase’ (RP) in
117 autumn-winter, Hirche et al., 2019).

118

119 Shifts in copepod body size have the potential to impact upon both their prey and their predators.
120 Besides temperature, food concentration and predation pressure can also impact growth and
121 development rates, and the body size achieved by copepod adults (Hall et al., 1976; Abrams and
122 Rowe, 1996; Bonecker et al., 2011). Indeed, seasonal body size variation is influenced by food
123 availability not only in copepods (Durbin et al., 1983) but other invertebrates too (Berrigan and
124 Charnov, 1994). Other factors that relate to the strength of T-S response in copepods include the
125 feeding mode and taxonomic affiliation of the species (Horne et al., 2016).

126

127 Pelagic copepod species play a pivotal role as grazers of microplankton, and as conduits of
128 energy and material to higher trophic levels, including fish and predatory invertebrates (Banse,
129 1995; Beaugrand et al., 2003). It is still unclear how their seasonal body size distribution varies
130 from year to year, and among species, and how this may change with respect to climate
131 warming. Sea surface temperature has increased significantly worldwide in the last 100 years
132 (IPCC report, 2014). However, climate warming occurs with great inter-annual fluctuations, yet
133 few previous studies have investigated the effects of copepod body size over multiple years
134 (although see e.g. Evans, 1981) and no study has simultaneously considered the T-S response of
135 many species over several years at a single location. To achieve such multi-year, within-site
136 synthesis of species body size responses across a copepod community we investigated seven
137 contrasting copepod species (*Acartia clausi*, *Oithona similis*, *Paracalanus parvus*,
138 *Pseudocalanus elongatus*, *Calanus helgolandicus*, *Centropages typicus*, *Temora longicornis*) in
139 13 years sampled from a 27-year record at the Plymouth L4 site, an inshore site in the Western
140 English Channel. To have several species measured over so many years allows us to explore both

141 inter-annual changes in body size and whether such patterns could be related to drivers related to
142 climate change.

143

144 Previous interpretations of size responses of different species based on meta-analyses across
145 different geographical locations and different decades (e.g. Horne et al., 2016) are potentially
146 complicated by the confounding effects of location-specific environmental factors. These
147 confounding effects make it hard to tease out the various biotic and abiotic factors that shape the
148 T-S response. For example, in addition to bottom-up factors, zooplankton community density has
149 already been shown to be related to increased mortality of copepod early-stages and to decreased
150 population growth, egg production and hatching success rate (Ohman and Hirche, 2001; Franco
151 et al., 2017). However, very little is known about the influence of community abundance *in situ*
152 on individual copepod body size. We used, for the first time, an extensive dataset from a single
153 location across multiple years to address the following questions: 1) To what degree do the
154 seasonal temperature-size responses of the seven copepod species differ? 2) Do food
155 concentration, long-term temperature anomalies, or both, regulate copepod size changes across
156 years? 3) Are copepod body size and strength of T-S response also governed by factors that
157 depend on zooplankton community density?

158

159

160 **Methods**

161 All the samples used for this study were taken from the zooplankton time series at Plymouth

162 Station L4 ($50^{\circ}15'N$ $4^{\circ}13'W$), 13 km SSW of Plymouth, which has been sampled at weekly
163 resolution, weather permitting, since March 1988 by Plymouth Marine Laboratory (Harris, 2010;
164 <http://www.westernchannelobservatory.org.uk/>). The water is ~54 m deep, which stratifies
165 seasonally, with a thermocline forming typically from about May to September. Surface water
166 temperature typically ranges from about $9^{\circ}C$ in winter to $17^{\circ}C$ in summer. The study location is
167 affected periodically by freshwater outflow from the rivers Plym and Tamar at Plymouth (Rees
168 et al., 2009). Further details about the Plymouth L4 station and microplankton and zooplankton
169 sampling are reported respectively in Widdicombe et al., (2010) and Atkinson et al.
170 (2015).

171

172 **Environmental data**

173 An electronic probe for conductivity, temperature and depth analysis (CTD) is currently used at
174 L4, but sea surface temperature (SST) has been measured consistently throughout the whole time
175 series using a mercury-in-glass thermometer inside a stainless steel bucket of water collected at
176 the surface. For consistency across all our sampling years, we used these latter measurements,
177 which showed close agreements with surface values from the CTD (Atkinson et al. 2015).

178

179 **Food data**

180 Microplankton, which are used to represent the copepod food variable in our study, have been
181 routinely collected and identified on a near-weekly basis since October 1992. Paired water

182 samples collected from 10m depth were preserved at sea with acid Lugol's iodine and neutral
183 formaldehyde. Samples were analysed at Plymouth Marine Laboratory following the British and
184 European Standard EN 15204:2006 "Water quality - Guidance standard on the enumeration of
185 phytoplankton using inverted microscopy" (Utermöhl technique, more microplankton sampling
186 details in Widdicombe et al., 2010). For this study we used the carbon mass density values
187 (measured as mg C m⁻³) of diatoms, dinoflagellates, coccolithophorids and *Phaeocystis*, as well
188 as ciliates and flagellates, from the following copepod data years that had microplankton data
189 availability: 1993, 1997, 2000, 2003, 2005, 2008, 2010, 2012, 2013 and 2014. The flagellate
190 group includes taxa that are nanoplanktonic in size, although only the larger nanoflagellates were
191 counted using the above method. However for simplicity we term our copepod food variable
192 "microplankton" since this group encompasses the size range which the copepods can eat.

193

194 **Zooplankton data**

195 Zooplankton have been sampled since 1988. Each sample consisted of two replicate tows made
196 with a WP-2 net (56 cm diameter, 200 µm mesh) towed vertically from a depth of 50 m to the
197 surface. Each plankton sample was fixed in 4% buffered formalin. Laboratory analysis involved
198 counting and identifying large and/or rare taxa in a fraction of the original sample (a half, a
199 quarter or an eighth, depending on the abundance of the original), sub-sampled using a Folsom
200 plankton splitter (Atkinson et al., 2015; Highfield et al., 2010). Each replicate was given an
201 estimated abundance average, measured as individuals m⁻³ (John et al., 2001). The source of the
202 dataset used here can be found in Atkinson et al. (2019).

203

204 For this study, one sample per month from April to October was considered from each of the
205 years 1989, 1991, 1993, 1995, 1997, 2000, 2003, 2005, 2008, 2010, 2013, 2014, 2016. These
206 years were chosen to cover a wide range of temperature conditions between years at L4, with
207 1991 being the coldest, and 2014 the warmest year. Other years, with intermediate temperature,
208 were chosen in order to provide a wide range of coverage and conditions.

209

210 **Body size measurements**

211 One quarter of each plankton sample was removed and observed under the stereomicroscope in
212 order to identify the first 30 (when possible) adult females of each of the species: *Acartia clausi*,
213 *Oithona similis*, *Paracalanus parvus*, *Pseudocalanus elongatus*, *Calanus helgolandicus*,
214 *Centropages typicus* and *Temora longicornis*. In the case of *C. helgolandicus* a small number of
215 the congener *Calanus finmarchicus* may have been inadvertently included, as these occur rarely
216 at the site (a median composition of 4% of *C. finmarchicus*, according to Maud et al., 2015).
217 Each individual was photographed with a stereomicroscope digital camera and then processed
218 using the software ImageJ to determine prosome lengths, by measuring from the tip of the
219 cephalothorax to the end of the last pedigerous somite. To avoid erroneous size estimations of *C.*
220 *typicus* resulting from broken spines, the end of the prosome was considered not as the tips of its
221 two most posterior lateral spines, but as the mid-point where the prosome articulates with the
222 urosome. In *O. similis* the end of the prosome is very hard to identify because of its complex
223 structure and reduced size, so, instead, we measured to the connection between the first and the
224 second segment of the urosome, as this is more visible (as done in Sabatini and Kiørboe 1994).

225 Prosome lengths were converted into estimated carbon mass (μg) using appropriate species-
226 specific equations from the published literature (See Table S1). These equations were obtained
227 from copepods commonly sampled in the North Sea. However, for all analyses we tested both
228 estimated carbon mass and prosome length, in order to see if different patterns would emerge.

229

230 **Determination of temperature-dependence of body size reduction**

231 We defined the TDP for each species as between the month with the largest average size (1st
232 generation of copepods of the year) and the month with the smallest average size (last generation
233 of the year). To quantify the relationship within the TDP between mean adult body size in each
234 month and temperature in the previous 30 days, the mean carbon mass values for each species in
235 each month in each year was regressed against mean temperature recorded in the 30 days prior to
236 and including the sample day. We chose 30 days partly to correspond broadly with previous
237 studies on the generation times of the same seven species (Sabatini and Kiørboe, 1994; Irigoien
238 and Harris, 2003; Leandro et al., 2006; Carlotti et al., 2007; Stegert et al., 2007; Dzierzbicka-
239 Głowacka et al., 2011). For fitting the body size-temperature regressions ($\mu\text{g C individual}^{-1}$ vs.
240 $^{\circ}\text{C}$) we used exponential models as in Horne et al. (2016). The slopes from these exponential
241 regressions were transformed into percentage change in carbon mass per $^{\circ}\text{C}$ as: $(\exp^{\text{slope}} - 1) \times 100$
242 (Forster et al., 2012). These values have a negative value when mass decreases with warming,
243 and a positive value when body mass increases with increasing temperature.

244

245 To assess interannual change in body mass, carbon mass values were transformed into anomaly
246 values i.e., the observed value for a specific month in a given year, minus the mean value for that
247 month across all 13 years. Anomalies of temperature and food biomass concentrations were also
248 determined, but to ascribe these values to the copepod juvenile developmental period they were
249 calculated as mean values over the 30 days prior to the date at which the copepod body sizes
250 were determined. To assess the seasonal effect of food (estimated as the sum of diatoms,
251 dinoflagellates, coccolithophorids, *Phaeocystis*, flagellates and ciliates biomass (mg C m^{-3})) on
252 body size, the body mass values ($\mu\text{g C}$) were plotted against the average concentration of food in
253 the 30 days prior to the sample day. A multiple linear regression model was used to test the
254 effect of food and temperature together ($\log_{10}(\text{Carbon mass}) \sim \text{Temperature} + \text{Food}$). Food data
255 were derived from fewer years (nine years), so the relationships among food, temperature and
256 copepod body size were examined using the same nine sample years across all months.

257

258 To test the relationship between zooplankton density and adult female copepod body size, we
259 first tested for each month the effect of total zooplankton density on carbon mass values of all
260 the 7 copepod species (scaled, centred and pooled together with the z-score method: (value –
261 mean) / standard deviation). The only significant size-density correlation was in April, which had
262 the strongest beta coefficient ($\beta = -0.52, p < 0.01, n = 81$). We therefore regressed body mass
263 values of all copepod species in April against the density (ind m^{-3}) of the different groups of
264 zooplankton: 1) gelatinous zooplankton (i.e. cnidarians and ctenophores), 2) holoplanktonic
265 crustaceans, 3) fish larvae, 4) non-gelatinous/non-crustacean holoplankton, and 5) meroplankton.
266 These were all estimated as the mean density in the 30 days prior to the date on which body size

267 was determined. The second variable (holoplanktonic crustacean density) was the strongest and
268 only significant one. We therefore performed species-specific linear regression between carbon
269 mass and holoplanktonic crustacean density, for each month. Another multiple regression was
270 performed to assess the effect of holoplanktonic crustacean density, temperature and food on
271 copepod body size in April. In all multiple regressions we checked the collinearity of each
272 independent variable (those with a Variance Inflation Factor greater than 5 were discarded, as
273 suggested in James et al., 2017).

274

275 **Results**

276 **Seasonal variation in adult body size**

277 A total of 7098 adult females were measured from the seven copepod species studied. The
278 duration of the Temperature Dependent Phase (TDP) differed slightly among the species: this
279 period was from April to October for *Acartia clausi*, *Oithona similis*, *Paracalanus parvus*,
280 *Pseudocalanus elongatus*, and from May to October for *Calanus helgolandicus*, *Centropages*
281 *typicus*, *Temora longicornis* (Figure 1). In each species, body mass showed a strong significant
282 negative correlation with temperature, being weakest in *Paracalanus parvus* ($R^2 = 0.25$),
283 compared to 0.48-0.82 for the other species; Figure 2). Copepod carbon mass variation was
284 always better explained by temperature than by food biomass (as judged by the outcomes of the
285 models in Table 1).

286

287 **Inter-annual variation in copepod adult body size**

288 Simple linear regressions between monthly adult carbon mass and year were never significant
289 (no evident increase or decrease overall over years) for any species. However, any potential body
290 size reduction driven by climate change is likely to be obscured by great inter-annual fluctuations
291 in warming at the L4 site over time (Atkinson et al., 2015), including a lack of correlation
292 between warming anomalies in one season versus another season. For instance, temperature
293 during summer at L4 is not significantly correlated with the temperature in the preceding spring
294 ($R^2 = 0.09$; $p = 0.08$, $n = 31$). Therefore, relatively colder summers may sometimes follow
295 relatively warmer springs and *vice versa*. For this reason, monthly anomaly values of carbon
296 mass, temperature and food biomass across years were used to assess long-term changes in these
297 variables and their co-dependence in each of the copepod species. Monthly carbon mass
298 anomalies were negatively related to monthly temperature anomalies for: *Acartia clausi* and
299 *Centropages typicus* as shown in Figure S1 (Suppl. Information), revealing the existence of an
300 inter-annual T-S response along with the seasonal one. No species showed significant
301 regressions between their body mass anomaly and food biomass anomaly.

302

303 In April, the only month that showed a significant relationship between body size of all species
304 pooled and total zooplankton density, effects of densities of different groups of zooplankton on
305 the overall copepod carbon mass (z-scores) showed that holoplanktonic crustacean density was
306 the strongest and only significant explanatory variable ($\beta = -0.69$, $p < 0.05$). Across the seven
307 copepod species and seven months (49 combinations) examined, significant correlations were
308 found between \log_{10} of adult body carbon mass of a given copepod species and \log_{10} crustacean

309 holoplankton density in eight cases: six were confined to April (all species except *A. clausi*) and
310 one in October (*Calanus helgolandicus*). All these seven body mass-holoplankton density
311 relationships were negative (Figure 3), accounting for between 47 and 80% of the variance in log
312 adult copepod body carbon mass in April, and 38% in October (Figure 3).

313

314

315 **Discussion**

316 **Seasonal T-S response and taxonomic differences**

317

318 This study represents one of the most comprehensive comparisons of seasonal and inter-annual
319 shifts in body size of multiple copepod species at a single site, based on 7098 individuals
320 measured. Indeed, this work allows the assessment of pelagic copepod body sizes from a single
321 location and hence with similar environmental conditions and annual temperature range (~ 10
322 °C), which makes the comparison across species much easier than in previous meta-analyses, in
323 which different species had experienced different environments (Horne et al., 2016). Firstly, we
324 assessed that the TDP of our species started around April or May (depending on the species) and
325 generally terminated around October, similar to what was shown in previous *in situ* analyses
326 (Hirst et al., 1999; Horne et al., 2016; Cornwell et al., 2018). During the rest of the year (the so-
327 called ‘Resting Phase’, set roughly between November and March) there appears to be little or
328 no recruitment to the adult stage (Djeghri et al., 2018), hence, adult size in the resting phase is a
329 legacy of conditions at the end of the previous TDP, possibly coupled with any size-dependent
330 adult mortality during the resting phase. The ‘Transition Phase’ then follows, usually in the first

331 months of Spring (March, April, May), when a mix of older but smaller individuals (from the
332 resting phase), and newly recruited, larger, individuals are found (Hirche et al., 2019).

333

334 Among our seven species, the range in T-S strength (based on monthly size and temperature
335 values, across the TDP over 13 years) was wide: strongest values were ~8% to ~10% change in
336 C mass $^{\circ}\text{C}^{-1}$ for *Acartia clausi*, *Centropages typicus*, *Pseudocalanus elongatus* and *Temora*
337 *longicornis*. This last species had the most intense T-S response (-10.15% change in C mass $^{\circ}\text{C}^{-1}$),
338 with body size in October roughly half that at the start of the TDP. This reduction of body mass
339 may be associated with a change in copepod fitness, feeding rates and food web connectivity
340 (Hansen et al., 1994; Rice et al., 2014). The weakest T-S response (-2.93% change in C mass $^{\circ}\text{C}^{-$
341 1), demonstrated in *Paracalanus parvus* (Order Calanoida), was considerably weaker than that of
342 the other Calanoida, while *Oithona similis*, the only non-calanoid, had the second weakest T-S
343 response (-3.33% change in C mass $^{\circ}\text{C}^{-1}$). These marked differences concur with the weak T-S
344 response in both *O. similis* and *P. parvus* observed by Horne et al., (2016). This latter species
345 also showed the most inconsistent T-S relationship, with the lowest R^2 (0.25, see Table 1).
346 However, the seasonal body length distribution of *P. parvus* observed by Riccardi and Mariotto
347 (2000) and Hirche et al. (2019) was similar to the one we have found. A difference in the T-S
348 response strengths between Cyclopoida and Calanoida species, although not statistically
349 supported here (possibly because of limited number of taxa), has previously been observed by
350 Horne et al. (2016), who suggested this may relate to their different feeding strategies (active
351 filtering in calanoids vs. ambush-feeding in cyclopoids) and the difference in body mass scaling
352 of physiological rates between these different feeding types.

353

354 **Effects of food concentration on copepod body size**

355 The models showed that, seasonally, temperature was always a stronger predictor of copepod
356 body size than was food biomass (Table 2). It is possible, however, that our use of measured
357 body lengths rather than actual masses or biovolumes may have led to our overlooking effects of
358 food concentration on changes in body condition, hence actual carbon mass values. As Klein
359 Breteler and Gonzalez (1988) found in laboratory experiments, food is likely to affect more the
360 “fatness” of a copepod than its prosome length. However, the direct C mass determinations made
361 from L4 (Fig. 6 in Pond et al. 1996) show that even for the larger lipid-storing species *Calanus*
362 *helgolandicus* females have a -8.9% change in $\mu\text{g C } ^\circ\text{C}^{-1}$ (95% C.I.: -11, -6.8) during the TDP.
363 This does not differ significantly from the estimates in our study based on prosome length-mass
364 conversion (-6.2%; 95% C.I.: -7.5, -4.9).

365

366 **Inter-annual body mass variation**

367 We tested species-specific size anomaly against temperature anomaly from each individual
368 month and year, to see how the copepod body size would react in the same month of another
369 year with anomalous temperature (Figure S1). Only two species (*Acartia clausi* and *Centropages*
370 *typicus*) showed a significant negative correlation between carbon mass anomaly and
371 temperature anomaly. This shows the existence of a long-term T-S response in at least these two
372 species. Although R^2 values were ≤ 0.2 , in evolutionary terms, even low effect size values may
373 still be relevant and significant. High scatter may also have been caused by the low number of
374 years (13), lag effects, food seasonality, and the high scatter of the SST - time relationship at L4

375 itself ($R^2 = 0.17$, Atkinson et al., 2015). Temperature increases are apparent in the English
376 Channel, although with a shallow slope (0.02° C increase per year, Atkinson et al., 2015) and
377 great inter-annual variability, making climate-change impacts on the plankton community more
378 challenging to detect. Moreover, body size in April was not modulated by temperature or food,
379 and as described in the next section, density-dependent effects became more important then.
380 When April samples were removed, two further species showed significant and negative long-
381 term T-S responses: *Pseudocalanus elongatus* and *Oithona similis*. Therefore, it is reasonable to
382 hypothesize a generalized long-term T-S response in most of the copepod species, consistent
383 with similar findings in *Acartia tonsa* and *A. hudsonica* by Rice et al., (2014) and in fish by
384 Daufresne et al. (2009). A long-term decrease in body size, driven by global warming, would
385 have major implications for food web structure and function, given the strong size-structuring
386 (Webb, 2012) and importance of copepods in these food webs (Ware and Thomson, 2005).

387

388 **Copepod T-S response strength relates to zooplankton density**

389 To date, very few studies (e.g. Cyr and Pace 1993) have explored how body size of copepods
390 may be impacted by the abundance of their own species or of the whole community, and this
391 topic has been historically underexamined in the marine environment. Our analysis of large
392 numbers of individuals, sampled through multiple seasons at a single site, showed a strong
393 relationship between copepod size and holoplankton density (in April), providing new evidence
394 that density-dependent controls on copepod body size can operate *in situ*.

395

396 Are the changes in copepod body size reflecting density-dependent changes in competition or
397 predation? The timing of the strong relationship with zooplankton density, being found only in
398 April, may provide some clues to the cause. In the spring months of March and April, food levels
399 were still low and increasing, whereas their grazers had already reached fairly high biomasses
400 (Fig. 4; Atkinson et al. 2015). The period preceding and including April thus experienced a low
401 food concentration to grazer biomass ratio (Fig. 4), pointing to the potential for increased
402 competition for food at this time. The ensuing competition for food may provide one explanation
403 for the reduced copepod size with increased community density in April.

404

405 An alternative or additional explanation for the density-dependent effects observed in April may
406 relate to direct predation. Work at L4 has increasingly emphasized the role of predation in
407 regulating copepod abundance and phenology (e.g.: Bonnet et al., 2005; Atkinson et al. 2015;
408 Cornwell et al 2018). Some of the food consumed by adult copepods is in fact often early stages
409 of copepods (Irigoinen and Harris, 2006; Boersma et al., 2014; Djeghri et al., 2018), as well as
410 eggs (Hirst et al., 2007; Maud, 2017; Cornwell et al., 2018). Ohman and Hirche (2001) suggested
411 that cannibalism could even be a form of population self-limitation from their study on *Calanus*
412 *finmarchicus*, and likewise Maud (2018) concluded the same for *C. helgolandicus* at L4. This
413 intraguild/cannibalistic predation on young stages would explain a reduced size at adulthood
414 according to Abrams and Rowe (1996), who affirm that high mortality rates on early stages
415 favour earlier maturity (hence smaller size at adult). The same phenomenon occurs for fish when
416 exposed to intense fishery activity (Olsen et al., 2004). However, this would imply that juvenile
417 copepods must produce an adaptive plastic response (developmental acceleration) to predation
418 cues during the Transition Phase (April), when exposed to high predation risk. Hence, the

419 intraguild predation hypothesis seems to need several assumptions in order to be sustained,
420 whereas the competition for food appears a simpler explanation. As is apparent in Figure 4,
421 increases in zooplankton abundance at L4 tend to occur well before the spring bloom, a
422 phenological effect seen across a suite of the component taxa including copepods (Atkinson et al.
423 2015). It is possible that the first cohorts of copepods of the TDP (around April) that are highly
424 abundant, when food is relatively scarce, will suffer from a poor diet and reach adulthood at a
425 more reduced size than usual (Durbin and Durbin, 1978; Klein Breteler and Gonzalez, 1988).

426

427 In this study we observed that the T-S response strength varies substantially among different
428 copepod species. This variation may relate to different thermal sensitivities of mechanisms
429 generating thermal reaction norms for growth and development rates (Verberk et al. 2020).
430 Intriguingly, large differences between species also exist in the two other “universal” responses
431 of copepods to climate change: the phenological response (Atkinson et al., 2015; Mackas et al.,
432 2012) and the distributional change (Chivers et al., 2017). These contrasting responses are also
433 manifested in their long-term population trajectories of the major copepod species. Most have
434 undergone substantial declines in summertime abundance across the NE Atlantic, whereas a few
435 species have increased (Schmidt et al. 2020). It is therefore an urgent priority to explore whether
436 these responses to climate change are inter-related, paying particular attention not just to bottom-
437 up controls, but also to top-down factors.

438

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440

441

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687

688 **Acknowledgments**

689

690 First, we thank the whole Plymouth Marine Laboratory community for welcoming and hosting S.
691 Corona during his copepod body size analysis. We thank everyone who contributed to the
692 sampling and analysis of plankton and environmental data at Station L4, from 1988 to today,
693 along with everyone who will keep doing so in future. In particular, we want to thank Andrea
694 McEvoy for training and helping S. Corona with the challenging task of copepod identification,
695 with proper methods of preparation and conservation of delicate zooplankton samples. Last but
696 not least we deeply thank the reviewers and editors for their helpful contribution in the
697 improvement of our paper. A. Atkinson's contribution was funded by the UK Natural
698 Environment Research Council (NERC) through its National Capability Long-term Single
699 Centre Science Programme, Climate Linked Atlantic Sector Science, grant number
700 NE/R015953/1, contributing to Theme 3.1—Biological dynamics in a changing Atlantic and 4—
701 Fixed Point Observations (Western Channel Observatory). The University of Liverpool funded
702 the Postgraduate Research Studentship of S. Corona.

703

704 **Conflict of interest**

705 None declared

706 **Tables and Figures**

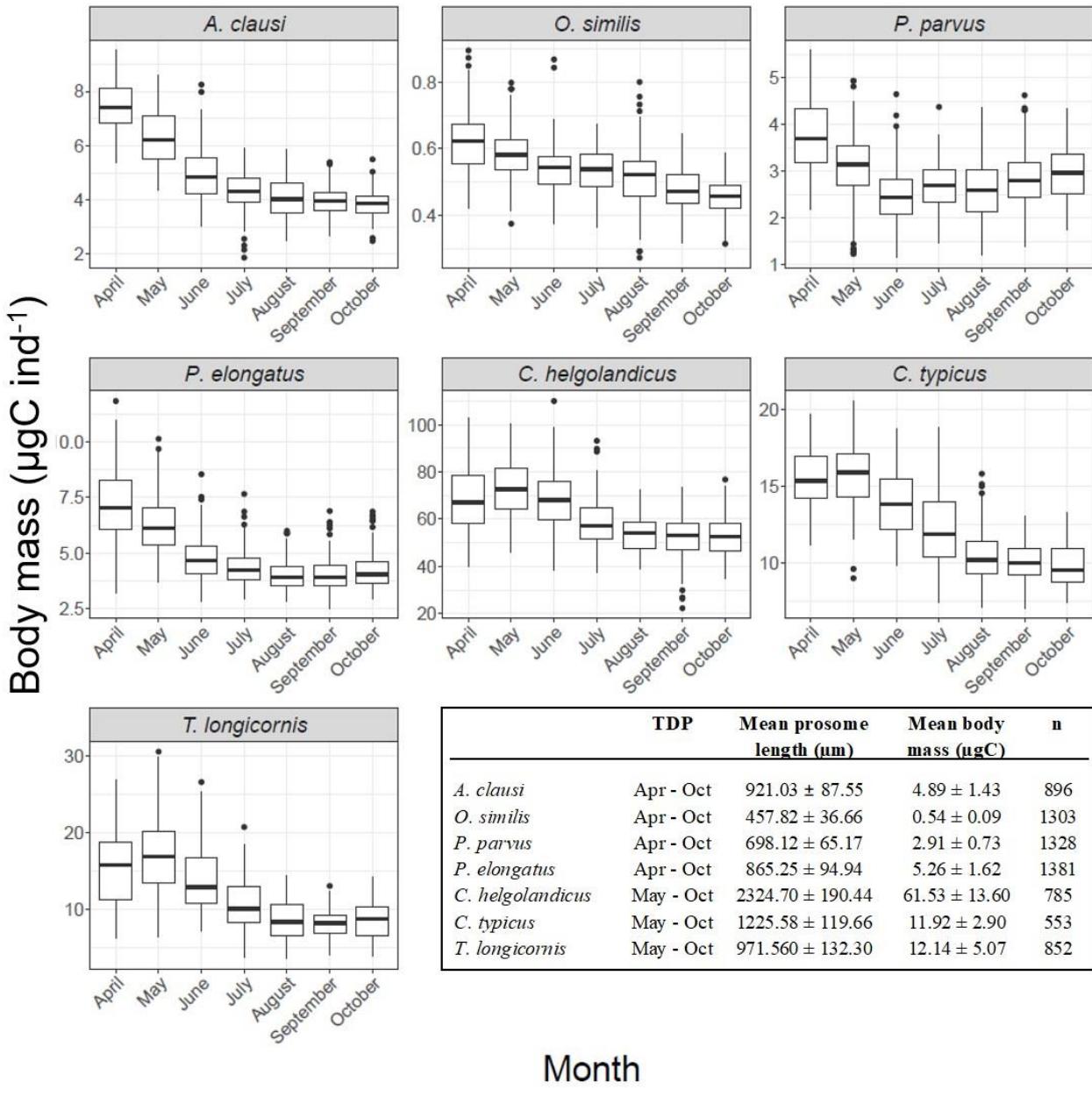
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708 Table 1: Multiple linear regressions coefficients: Standardized beta coefficients (β_T refers to
 709 temperature coefficient and β_F refers to food biomass coefficient, in bold when significant) R^2
 710 and p values of each model. The regression equations with actual units for each species are here
 711 reported: *A. clausi*: $\log_{10}M = 2.75 - 0.08T + 3 \times 10^{-4}F$; *P. parvus*: $\log_{10}M = 1.74 - 0.02T - 0.07F$;
 712 *P. elongatus*: $\log_{10}M = 2.83 - 0.07T - 0.05F$; *O. similis*: $\log_{10}M = -0.30 - 0.04T + 0.05F$; *C.
 713 helgolandicus*: $\log_{10}M = 4.61 - 0.05T + 0.06F$; *C. typicus*: $\log_{10}M = 3.26 - 0.07T + 0.07F$; *T.
 714 longicornis*: $\log_{10}M = 3.70 - 0.09T + 0.02F$ (note that M refers to body mass measured as $\mu\text{g C}$,
 715 T refers to temperature measured as the average $^\circ\text{C}$ of sea surface in the 30 days before sampling
 716 day, and F refers to food biomass measured as the average mg C m^{-3} in the 30 days before
 717 sampling day).

718

RESPONSE PREDICTOR	Log ₁₀ (Body mass)			
	Temperature + Food biomass			
	β_T	β_F	R^2	p
<i>Acartia clausi</i>	-0.90	0.00	0.81	< 0.01
<i>Paracalanus parvus</i>	-0.41	-0.26	0.30	< 0.01
<i>Pseudocalanus elongatus</i>	-0.79	-0.14	0.73	< 0.01
<i>Oithona similis</i>	-0.82	0.27	0.54	< 0.01
<i>Calanus helgolandicus</i>	-0.77	0.18	0.48	< 0.01
<i>Centropages typicus</i>	-0.90	0.18	0.74	< 0.01
<i>Temora longicornis</i>	-0.76	0.04	0.54	< 0.01
Average	-0.76	0.04		

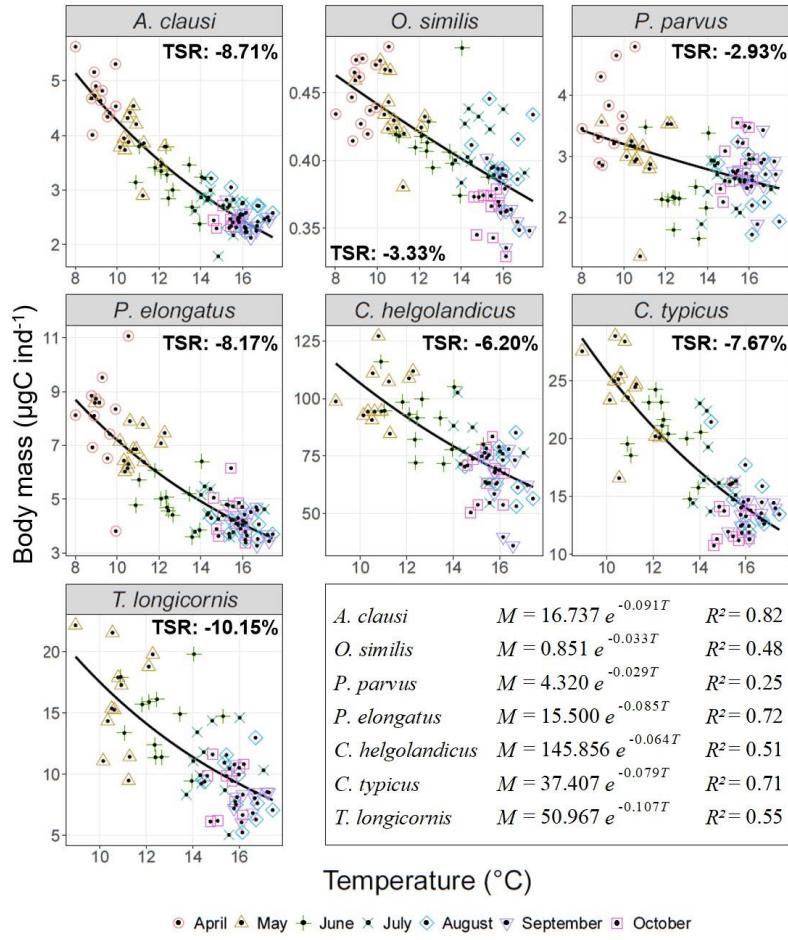
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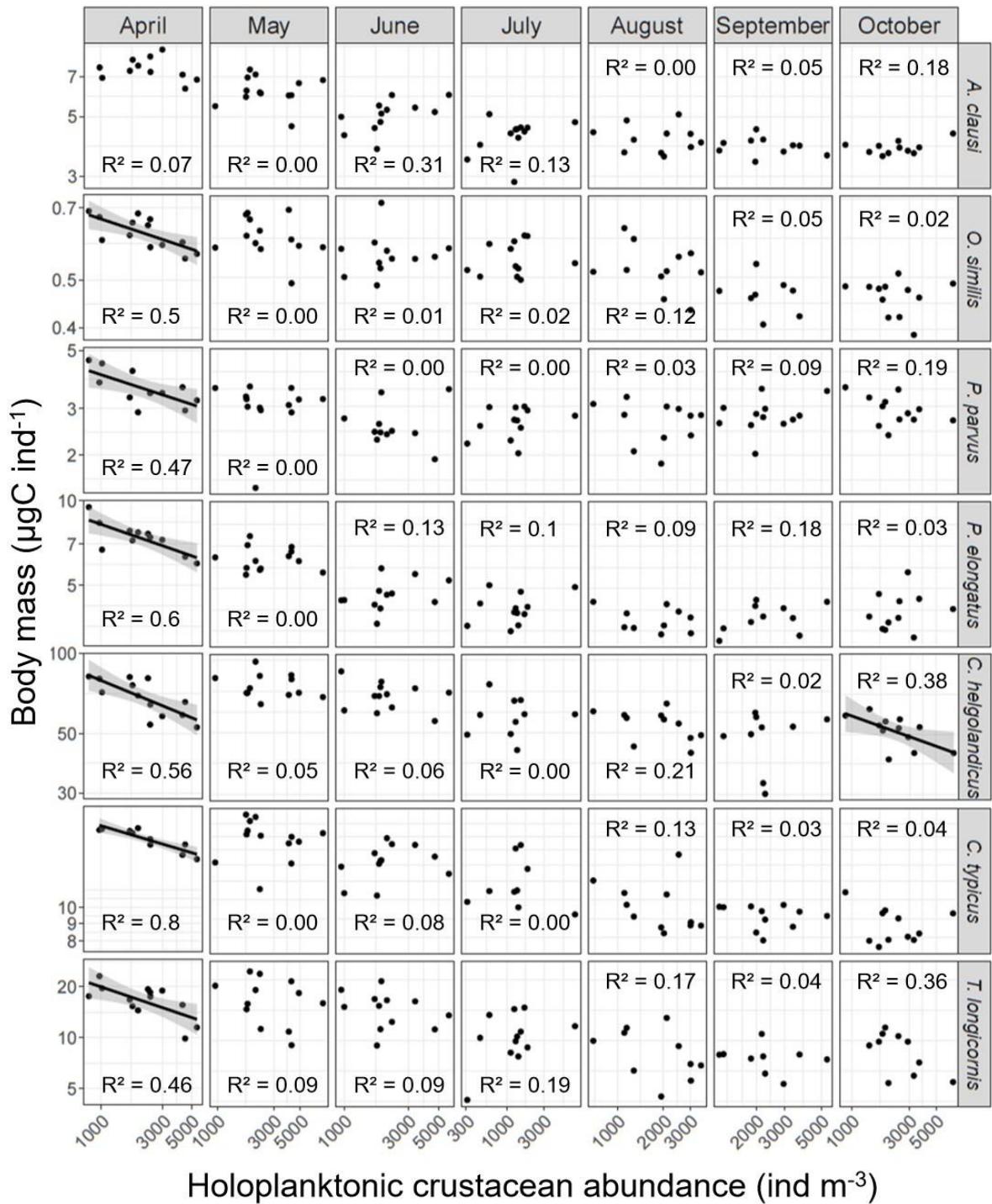
721 Figure 1: Seasonal body mass distribution ($\mu\text{g C ind}^{-1}$) of adult females for each of the seven
 722 pelagic copepod species at the L4 site. The TDP of each species is reported in the box on the
 723 bottom-right, along with species-specific mean prosome length and mean body mass values (both
 724 with corresponding \pm standard deviations) across all samples.

725



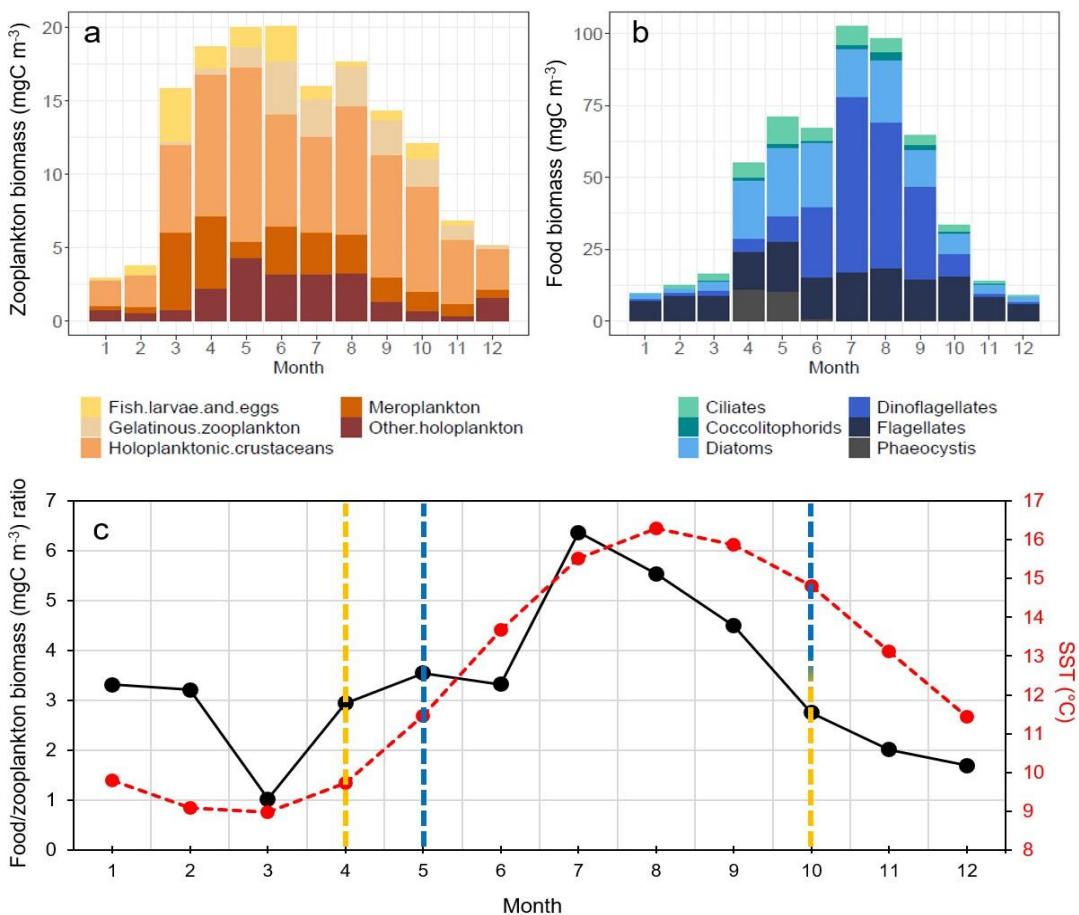
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727 Figure 2: Exponential regressions (solid lines) between monthly average body mass ($\mu\text{g C ind}^{-1}$)
 728 values and temperature ($^{\circ}\text{C}$), across all 13 years studies. On each plot the TSR is also reported
 729 (expressed as % change in $\mu\text{g C }^{\circ}\text{C}^{-1}$). Values for each month are indicated by the different
 730 symbols, see key in the lowest part of the figure. In the box on the bottom-right, each exponential
 731 regression is reported along with its corresponding R^2 value (note: M = body mass, T =
 732 temperature.



733

734 Figure 3: Monthly mean body mass of the adult females of the 7 copepod species ($\mu\text{g C ind}^{-1}$) plotted
 735 against holoplanktonic crustacean abundance (ind m^{-3}). Associated R^2 values are reported, solid
 736 black lines represent significant regressions. Note that both the x and y scales are on a \log_{10} scale.



737

738 Figure 4: a) Zooplankton biomass (mg C m^{-3}) distribution across months for five major functional
 739 groups (see key; note that “holoplanktonic crustaceans” is the group whose density in April is
 740 inversely correlated with copepod body mass in April (Figure 3); “other holoplankton” refers to non-
 741 crustacean and non-gelatinous holoplankton); b) food (microplankton) biomass (mg C m^{-3}) across
 742 months for six major functional groups (see key); c) food/zooplankton biomass ratio (black solid
 743 lines) and sea surface temperature (red dashed lines) across months, vertical dashed lines indicate the
 744 boundaries of the TDP April- October (yellow vertical line) and the TDP May-October (blue vertical
 line). Plots are based on the whole L4 time-series (1988-2019).