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Freshwater Biology

Faunal response to benthic and hyporheic sedimentation varies with direction of vertical hydrological exchange.

Journal:	<i>Freshwater Biology</i>
Manuscript ID:	FWB-A-Jan-14-0060
Manuscript Type:	Applied Issue
Date Submitted by the Author:	28-Jan-2014
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Keywords:	Erosion / sedimentation / landuse < Applied Issues, Hyporheic zone < Habitat, Population < Level of Organisation, Experimental ecology < Process / Approach / Methods, Invertebrates < Taxonomic Group / Assemblage

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6 3 **of vertical hydrological exchange.**
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4 26 **of vertical hydrological exchange.**
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8 28 Mathers, K.L.¹ Millett, J.¹ Robertson, A.L.² Stubbington, R.³ and Wood, P.J.¹
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11
12 30 **Abstract**
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15 31 1. Sedimentation and clogging of benthic and hyporheic zone substrates is
16 32 increasingly being recognised as one of the greatest threats to the ecological
17 33 integrity of riverine ecosystems globally. This *ex-situ* study examined the
18 34 influence of sedimentation (benthic and hyporheic) and pattern of hydrological
19 35 exchange on the vertical distribution of the freshwater shrimp *Gammarus*
20 36 *pulex* within the experimental substrates of running water mesocosms.
21 37
22 38 2. Six sediment treatments representing a continuum from a clean gravel
23 39 substratum through to heavy sediment loading of both benthic and hyporheic
24 40 substrates were used to examine the distribution of *G. pulex* in relation to the
25 41 direction of hydrological exchange (downwelling, upwelling and no exchange).
26 42
27 43 3. The vertical distribution of fauna varied significantly for both sediment
28 44 treatment and pattern of hydrological exchange. There was a significant
29 45 interaction between the two effects indicating that the effect of sedimentation
30 46 varied depending on the pattern of vertical hydrological exchange.
31 47
32 48 4. Sedimentation of benthic sediments resulted in significant modification to the
33 49 distribution of *G. pulex* when there was no hydrological exchange (no flow)
34 50 within the column, although there were only limited changes with downwelling
35 51 flow and no statistical differences with upwelling flow.
36 52
37 53 5. Sedimentation of multiple layers of the column (benthic and hyporheic)
38 54 significantly reduced the ability of individuals to utilise the lower layers of the
39 55 substratum (i.e. the hyporheic zone). This was most marked for upwelling
40 56 conditions, where it resulted in a complete reversal of the vertical distribution
41 57 pattern recorded.
42 58
43 59 6. This study demonstrates that faunal movement, and use of benthic and
44 60 hyporheic substrates, may be influenced by sedimentation and modified by
45 the pattern of vertical hydrological exchange. Severe sedimentation
46 (colmation) has the potential to prevent benthic fauna from accessing the
47 hyporheic zone and its resources.
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51 **Keywords:** surface water and groundwater exchange, upwelling and downwelling,
52 fine sediment, colmation, *Gammarus pulex*, mesocosm experiment.
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63 Introduction

64 Fine sediment deposition (sedimentation) is widely implicated as a major contributor
65 to ecosystem impairment across the globe (Walling, 2009; Collins et al., 2011).
66 Changes in agricultural practices and land use (Stundinski et al., 2012), urbanisation
67 (Taylor & Owens, 2009; Wang et al., 2012) and channel management / habitat
68 modification (Dunbar et al., 2010) have increased the erosion and delivery of fine
69 sediments (typically referred to as particles <2 mm in diameter) to aquatic
70 ecosystems. The effects of excessive sedimentation have been demonstrated at all
71 trophic levels from fish (Walters et al., 2003; Kemp et al., 2011) and benthic
72 invertebrates (Larson & Ormerod, 2010; Jones et al. 2012a), through to macrophytes
73 and periphyton (Izagirre et al., 2009; Jones et al., 2012b), although the potential
74 effects on biota within the hyporheic zone are less well defined (Boulton et al., 2010;
75 Richards & Bacon, 1994).

76 The hyporheic zone is a dynamic ecotone situated below the river bed which is
77 composed of saturated sediments that exchange water between the surface stream
78 and underlying groundwater (White, 1993). The hyporheic zone represents the
79 interface between the river channel and groundwater and plays a key role in many
80 hydrological and biogeochemical processes in river systems (Boulton & Hancock,
81 2006; Robertson & Wood, 2010). Consequently, the hyporheic zone is increasingly
82 being recognised as an integral component of lotic ecosystems (Krause et al., 2011)
83 with vertical hydrological connectivity now widely recognised as being a strong
84 determinant of the patterns observed (Malard et al., 2002; Boulton, 2007; Heppell et
85 al., 2009).

86 Hydrological exchange between surface water and groundwater occurs at a variety
87 of spatial and temporal scales, resulting in a mosaic of habitat patches which are
88 characterised by differing connectivity, permeability and physio-chemical conditions
89 (Dole-Oliver & Marmonier, 1992; Krause et al., 2011). Surface water enters the river
90 bed and hyporheic zone when the hydraulic head is greater than that of the
91 groundwater (downwelling water). This water may be subject to further exchanges,
92 either passing deeper into the groundwater zone or travelling through the sediments
93 until the water emerges from the interstices (upwelling water) (Brunke & Gonser,
94 1997; Krause et al., 2011). Alternatively, in some rivers there may be limited or no

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3 95 vertical hydraulic exchange due to limited connectivity between surface and
4 96 groundwater or the presence of layers of impermeable bedrock (Ryan & Boufadel,
5 97 2006; Malcolm et al., 2003).

8 98 The pattern of vertical hydrological exchange is one of the primary controls of
9 99 dissolved oxygen concentrations (Olsen and Townsend, 2003), thermal
10 100 characteristics (Evans & Petts, 1997), nutrient levels (Franken et al., 2001), stream
11 101 metabolic processes and organic matter breakdown in alluvial rivers (Krause et al.,
12 102 2011). Patterns of hydrological exchange have also been shown to be associated
13 103 with distinct benthic (Pepin & Hauer, 2002; Davy-Bowker et al., 2006) and hyporheic
14 104 invertebrate communities (Plenet et al., 1995; Fowler & Scarsbrook, 2002).

21 105 Interstitial sedimentation has the potential to reduce the porosity and permeability of
22 106 the substratum (Boulton et al., 1998; Bo et al., 2007), thereby limiting the vertical
23 107 exchange of water and nutrients across the surface and groundwater ecotone
24 108 (Brunke, 1999; Descloux et al., 2013). The associated reduction of available
25 109 interstitial habitat may also lower hyporheic metabolism and productivity, and
26 110 significantly reduce the ability of fauna to exploit the resources of the hyporheic zone
27 111 (Boulton, 2007; Descloux et al., 2013). The accumulation of fine sediments is not
28 112 ubiquitous, and patterns of sediment deposition and erosion vary temporally
29 113 reflecting the flow regime, fine sediment availability and local channel morphology
30 114 (Boano et al., 2007). As a result, gravel-bed rivers are commonly comprised of a
31 115 mosaic of substratum patches which are characterised by variable patterns of
32 116 vertical hydrological exchange, and fine sediment deposition and flushing processes
33 117 (Dole-Olive and Marmonier, 1992; Boulton & Stanley, 1995).

44 118 Fine sediment particles may be deposited on the surface of the river bed (benthic
45 119 sedimentation) or transported into the river bed where they may be deposited
46 120 beneath the armour layer or deeper in the hyporheic zone (Huettel et al., 1996; Ren
47 121 & Packman, 2007; Simpson & Meixner, 2012). This clogging of the interstices
48 122 directly below the armour layer is typically referred to as colmation, and it may result
49 123 in the formation of a thin seal (clog) which can disconnect surface water and
50 124 hyporheic habitats (Brunke, 1999; Packman & Mackay, 2003). Consequently,
51 125 colmation of hyporheic sediments may be present even in the absence of benthic
52 126 fine sediment deposits, and surface sedimentation may be poorly correlated with

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3 127 subsurface colmation (Descloux et al., 2010). Ultimately, high levels of fine sediment
4 128 deposition may lead to the filling of interstitial spaces, particularly in areas of
5 129 downwelling water (Brunke & Gosner, 1997). In contrast, strongly upwelling water
6 130 may maintain open pathways of flow by preventing further fine sediment ingress and
7 131 in some instances flushing fines from the bed (Packman & Salehin, 2003).

11 132 The effects of sedimentation on the hyporheic zone are now regarded as a
12 133 significant ecological threat to many rivers (Boulton, 2007; Boulton et al., 2010). To-
13 134 date published studies assessing sedimentation effects on hyporheic faunal
14 135 communities have reported a reduction in the density and/or diversity of fauna with
15 136 increasing volumes of fine sediment (Bo et al., 2007; Sarriquet et al., 2007; Bruno et
16 137 al., 2009; Pacioglu et al., 2012; Descloux et al., 2013; 2014). The relatively small
17 138 number of studies reflects the difficulties associated with replicating and quantifying
18 139 natural hyporheic fine sediment concentrations prior to the onset of experimental
19 140 conditions within spatially heterogeneous alluvial river beds (Descloux et al., 2013).
20 141 As a result, the need for controlled and replicated *ex-situ* experimental approaches is
21 142 increasingly being recognised in groundwater ecology (Stump & Hose 2013; Navel et
22 143 al., 2012; Larned, 2012).

23 144 In this *ex-situ* study, the vertical distribution of the freshwater amphipod, *Gammarus*
24 145 *pulex* (L.) (Amphipoda: Crustacea) was examined in response to different patterns of
25 146 vertical hydrological exchange and sedimentation (benthic and hyporheic) within
26 147 experimental running water mesocosms. *G. pulex* is a widespread and abundant
27 148 model organism that has been extensively studied (Sutcliffe, 1993). It is known to
28 149 colonize benthic, hyporheic and hypogean habitats within the UK (Gledhill et al.,
29 150 1993). In many riverine communities *G. pulex* is the dominant macroinvertebrate in
30 151 terms of biomass (MacNeil et al., 1997). It is moderately sensitive to fine sediment
31 152 and is capable of burrowing through substrate to find suitable habitat / resources
32 153 (Sutcliffe, 1993; Extence et al., 2013). Consequently, any alterations to the
33 154 distribution of this taxa are likely to represent the effect of sedimentation on the wider
34 155 invertebrate community.

35 156 We hypothesised that the vertical distribution of *G. pulex* would be influenced by
36 157 sedimentation and vertical hydrological exchange. Specifically, we predicted that: i)
37 158 increasing levels of fine sedimentation would modify the vertical distribution of *G.*

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3 159 *pulex* within the experimental columns by limiting and/or preventing movement in to
4 deeper sections; and ii) the influence of sedimentation on the vertical distribution of
5 individuals within the columns would differ for each pattern of hydrological exchange
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8 162 (no exchange downwelling or upwelling).
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10 163 **Methods**

11 164 *Experimental sediment columns*

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14 165 Experiments were undertaken within two identical sediment columns of five
15 interlocking sections / layers (Figure 1 – sections A-E). Each section was 22 cm in
16 166 diameter and contained 50 mm depth of coarse riverine sediments (gravel particles
17 167 20-64 mm in diameter). The sections stacked vertically to provide a total sediment
18 168 column depth of 250 mm. Ten holes (10 mm diameter) were drilled into the base of
19 169 four sections (0 - 200 mm depth) to allow water and organisms to pass between
20 170 sections. The final section (200-250 mm depth) was perforated with smaller holes (2
21 171 mm diameter) to permit the vertical exchange of water but prevented the movement
22 172 of individuals outside of the experimental column. In addition, 0.25 mm mesh sieves
23 173 were placed over the base and the top of the sediment columns for the duration of
24 174 each experiment, and a 5 mm rubber seal was created around the base of each
25 175 section to prevent the migration of individuals outside the column.
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29 177 The sediment columns were placed inside separate large cylindrical black plastic
30 178 water containers (90x40 cm, volume = 100 L). Two external pumps delivered flowing
31 179 water to the columns (4.5-4.8 L min⁻¹). This flow of water was sufficient to maintain
32 180 low interstitial flow velocity through the sediments but was not high enough to
33 181 transport or erode the deposited sediments. Consequently any movement of fine
34 182 sediments during the experimental period was primarily a function of gravity and
35 183 bioturbation associated with the movement of *G. pulex*. Three different hydrological
36 184 flows were simulated; no exchange, downwelling and upwelling.

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39 185 Downwelling conditions were simulated by pumping water directly into the top
40 186 experimental section and allowing water to pass through the column under gravity.
41 187 To mimic upwelling conditions, water was pumped through a large funnel / diffuser
42 188 (200 mm diameter) placed at the base of the column. Water rose through the column
43 189 and was allowed to overflow. The top of the column was covered with a 0.25mm
44 190 mesh to prevent *Gammarus* escaping. Both standing water (no exchange) and
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3 191 downwelling experiments were conducted with 10 cm depth of water over the
4 192 substratum to mimic overlying surface water. The experimental containers were
5 193 aerated through the use of an aquaria aeration pump and held at a constant
6 194 temperature (15°C +/- 0.4°C) via an external water-cooler (Aqua medic, Titan 150).
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10 195 Fine sediments used in the experiment comprised of pre-washed riverine sands
11 196 (0.125 µm - 1 mm in diameter). Silt and clay fractions (<0.125µm) were removed
12 197 through wet sieving to ensure that turbidity did not vary between experiments. Prior
13 198 to each experiment, fines were applied evenly to the surface of each wet gravel
14 199 section using a 1 mm sieve. Preliminary tests indicated that the application of an
15 200 equivalent of 5 kg m⁻² filled all interstices (100% of interstitial volume) of each section
16 201 and covered the surface of all gravel particles. Six fine sediment treatments were
17 202 examined (Figure 1) and for all treatments 50 mm of gravel was placed in each
18 203 section prior to fine sediment treatment:
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26 204 1. An open gravel framework: 50 mm depth of gravel in all sections of the column
27 205 (control);
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30 206 2. Benthic sedimentation: the equivalent of 3 kg m⁻² fine sediment applied to the top
31 207 section resulting in the clogging of 55-60% interstitial volume (0-50 mm – section A);
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34 208 3. Heavy benthic sedimentation: the equivalent of 5 kg m⁻² fine sediment applied to
35 209 the top section (section A);
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38 210 4. Hyporheic sedimentation of one section: the equivalent of 3 kg m⁻² fine sediment
39 211 applied to section C (100-150 mm depth);
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43 212 5. Hyporheic sedimentation of three sub-surface sections (simulating hyporheic
44 213 clogging): the equivalent of 3 kg m⁻² applied to sections B, C and D (50-100 mm,
45 214 100-150 mm and 150-200 mm); and
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49 215 6. Benthic and hyporheic-sedimentation (simulating benthic and hyporheic clogging)
50 216 – the equivalent of 3 kg m⁻² applied to all five layers (sections A, B, C, D and E).
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53 217 The sediment treatments (*n*=6) and patterns of hydrological exchange (*n*=3) were
54 218 combined in a full-factorial design giving 18 treatment combinations. Each
55 219 combination was replicated 6 times to give a total of 108 individual experiments.
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58 220 Treatments were randomly allocated to a run.
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3 221 All *G. pulex* specimens were collected from a local stream where the taxon occurs at
4 222 high abundances (>100 individual per m⁻²). Twenty-five individuals of mixed sizes
5 223 (5-16 mm length) were released onto the top section of the prepared column (0-50
6 224 mm) and left for 24-hours to allow individuals to redistribute within the sediment
7 225 columns. A single pre-conditioned horse chestnut leaf (*Aesculus hippocastanum*)
8 226 was placed in each section for food (Joyce et al., 2009). At the end of each
9 227 experiment, individuals were collected from each section by washing the contents of
10 228 each section through 0.25 mm sieves. All fine sediments were removed from the
11 229 column and retained for use in subsequent experimental runs.
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231 **Statistical Analysis**

232 Differences in the abundance of *G. pulex* in each section of the column relative to
233 the impact of sedimentation and pattern of vertical hydrological exchange were
234 tested using a Linear Mixed Model (LMM) in IBM SPSS Statistics 20.0 (IBM Corp.
235 2011). 'Section' was specified as a fixed within-subject (repeated) effect and the
236 pattern of hydrological exchange and sedimentation treatments were specified as
237 fixed between-subject effects. Covariance between sections of the columns was
238 modelled using a compound symmetry (CS) covariance structure. The model was
239 tested using an AR(1) covariance structure, but assessment of Akaike's information
240 Criterion (AIC) indicated that the CS covariance structure was more appropriate. The
241 model was fitted using Residual / Restricted Maximum Likelihood (REML) estimation.
242 Differences between sections within each treatment combination were tested using a
243 Fisher's LSD post-hoc test.
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245 **Results**

246 *Faunal response to the pattern of sedimentation and vertical hydrological exchange*

247 Overall recapture rates for all experiments were 91%, with downwelling and
248 upwelling treatments having the highest and lowest rates respectively (96% and
249 88%). When all experiments were considered *G. pulex* abundance was greatest in
250 the top column section (A) (mean \pm SE = 8.04 \pm 0.237), followed by the second (5.33
251 \pm 0.237) and the bottom (200-250 mm) layer (5.31 \pm 0.237). The fewest *G. pulex*
252 were recovered from the third (section C) and fourth (section D) layers of the
253 columns (2.31 \pm 0.237 and 1.82 \pm 0.237 respectively). However, the extent and
254 pattern of these differences was significantly affected by both the sedimentation

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3 255 treatment ($F_{20,360} = 13.05$, $P = <0.001$) and the pattern of vertical hydrological
4 256 exchange ($F_{2,90} = 7.43$, $P = <0.001$). There was a significant interaction between the
5 257 effect of these two treatments on the abundance of *G. pulex* in all sections ($F_{40,360} =$
6 258 6.27 , $P = <0.001$). As such, the effect of sedimentation on the distribution of
7 259 abundance within the sections of the columns differed depending on the pattern of
8 260 vertical hydrological exchange.
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14 262 *Faunal response to sedimentation under no exchange conditions.*

15 263 There were marked differences in the distribution of *G. pulex* when subjected to
16 264 varying levels of sedimentation under no-exchange conditions (Figure 2). In the open
17 265 gravel experiments, the greatest number of individuals was recorded in the top
18 266 (section A) and bottom (section E) layers of the column (Figure 2a). In the benthic
19 267 sedimentation treatments, a significantly higher number of individuals were recorded
20 268 in the second (section B) and bottom section (E) of the column for the moderate (3
21 269 kg m^{-2}) sedimentation treatment (Figure 2b), and in the second section for the heavy
22 270 (5 kg m^{-2}) sedimentation treatment (Figure 2c). Hyporheic sedimentation of the third
23 271 layer (section C) resulted in a similar pattern to that recorded for the heavy (5 kg m^{-2})
24 272 benthic sedimentation treatment (Figure 2d). Hyporheic sedimentation of 3 layers
25 273 (sections B, C and D) and all layers of the sediment column resulted in significantly
26 274 higher numbers of individuals in the top and second layers (section A and B) (Figure
27 275 2e and Figure 2f).
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40 277 *Faunal response to sedimentation under downwelling conditions.*

41 278 The distribution of *G. pulex* in all downwelling hydrological exchange experiments
42 279 was characterised by a reduction in the number of individuals with increasing depth
43 280 in the column (Figure 3). The majority of individuals were recorded in the top layer of
44 281 the column (section A) for the open gravel treatment, with <5 individuals typically
45 282 recorded in the lower sections (Figure 3a). When sedimentation of the benthic layer
46 283 occurred a significantly higher number of individuals were recorded in the top and
47 284 second layers (section A and B) for the moderate treatments (3 kg m^{-2}) and in the
48 285 top, second and third layers (section A-C) for heavy sedimentation (5 kg m^{-2})
49 286 treatments (Figure 3b and Figure 3c). Hyporheic sedimentation of one layer (section
50 287 C) resulted in a less marked gradient (Figure 3d). Hyporheic sedimentation of 3
51 288 layers (sections B, C and D) and all layers of the sediment column resulted in similar
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3 289 gradients with the majority of individuals being recorded in the top three layers
4 290 (sections A, B and C) and upper two layers (sections A and B) of the sediment
5 291 column respectively (Figure 3e and Figure 3f).
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9 293 *Faunal response to sedimentation under upwelling conditions.*

10 294 In upwelling experiments, *G. pulex* distribution was characterised by a significantly
11 295 greater number of individuals in the bottom section (section E) for control (Figure 4a),
12 296 moderate benthic sedimentation (Figure 4b), heavy benthic sedimentation (Figure
13 297 4c), and hyporheic sedimentation of one layer (section C) treatments (Figure 4d).
14 298 However, sedimentation of 3 hyporheic layers (sections B, C and D) resulted in no
15 299 statistical difference in the number of individuals among any sections of the column
16 300 (Figure 4e). Sedimentation of all layers of the column resulted in a complete reversal
17 301 the distribution of individuals compared to control conditions with a significantly
18 302 greater numbers being recorded in the top layer (section A) of the column (Figure 4f).
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28 304 **Discussion**

29 305 The results of the experiments presented in this study provide evidence to support
30 306 our first hypothesis (increasing levels of sedimentation will modify the vertical
31 307 distribution of *G. pulex* within the experimental columns). Sedimentation of the
32 308 benthic layer (0-50 mm depth) did not affect the distribution of individuals when
33 309 flowing water (upwelling or downwelling) occurred. Only in the absence of flow was
34 310 there a significant effect; although benthic sedimentation did not appear to impede
35 311 vertical movement as more individuals were recorded below the treated layer than
36 312 above it. In marked contrast, sedimentation of multiple layers (3 layers – 50-200 mm
37 313 depth and 5 layers – 0-250 mm depth) resulted in a significant reduction in the
38 314 abundance of individuals with increasing depth with the majority of individuals
39 315 confined to the top 100 mm of the substratum under the highest sediment loads;
40 316 although this did not modify the vertical distribution pattern of individuals during the
41 317 downwelling flow experiments. The deposition of fine sediments within riverine
42 318 substrates potentially reduces porosity and permeability (Ren & Packman, 2007;
43 319 Simpson & Meixner, 2012) leading to significant modification of interstitial habitat
44 320 characteristics. Sedimentation is widely reported to reduce benthic and hyporheic
45 321 interstitial habitat availability (Richards & Bacon, 1994; Gayraud & Philippe, 2003).
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3 323 The results of the experiments also provide evidence to support our second
4 324 hypothesis (the vertical distribution of individuals would differ for each of the patterns
5 325 of hydrological exchange - upwelling, downwelling and no exchange). There were
6 326 significant differences in the vertical distribution of taxa recorded for each of the
7 327 patterns of hydrological exchange when no fine sediment was present in the
8 328 columns. These differences persisted until fine sediment had been applied to
9 329 multiple layers of the substratum. When all benthic and subsurface layers were
10 330 treated with fine sediment the majority of individuals were recorded in the top and
11 331 second layers of the substratum under all hydrological conditions. For upwelling flow
12 332 conditions this represented a complete reversal in the pattern of vertical distribution
13 333 compared to control conditions and suggests that individuals were unable to migrate
14 334 through the column due to clogging of interstitial spaces. For the less severe fine
15 335 sediment treatments, pore space connectivity appears to have been maintained;
16 336 most clearly for the upwelling flow experiments.
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28 338 The results of field observations and experiments (Olsen & Townsend, 2003), and
29 339 theoretical insights (Krause et al., 2011) suggest that the effects of sedimentation on
30 340 macroinvertebrates may be modified by the nature of hydrological exchange.
31 341 *Gammarus pulex* is widely reported to be rheopolic, demonstrating a preference for
32 342 flowing water conditions (Gledhill et al., 1993). It was therefore not unexpected that
33 343 under control conditions (open gravel framework), the greatest number of individuals
34 344 were recorded in areas where the highest flow velocities occurred (in the benthic
35 345 zone for downwelling condition and at the base of the column for upwelling water).
36 346 Under no flow / hydrological exchange conditions and sedimentation of one section
37 347 of the substratum, the majority of individuals were recovered from the surface layers
38 348 (0-100 mm) or within the final section (200-250 mm). However, sedimentation of
39 349 multiple layers appeared to limit movement into the lower layers of the column (100-
40 350 250 mm).
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52 352 Fine sediment deposition (clogging / colmation) and the pattern of surface water –
53 353 groundwater exchange have been implicated as major factors in the structuring of
54 354 benthic and hyporheic faunal communities (Maridet et al., 1992; Richards & Bacon,
55 355 1994; Olsen & Townsend, 2003; Descloux et al., 2013). The ability of fauna to move
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3 356 and migrate from the surface stream (benthic zone) into the underlying groundwater
4 357 environment (hyporheic zone and groundwater) may be impeded in the presence of
5 358 heavy fine sediment loading (Boulton, 2007). Contrary to published reviews that
6 359 address the effects of benthic sedimentation on macroinvertebrates (see Wood &
7 360 Armitage, 1997; Jones et al., 2012a), heavy fine sediment loading of surface
8 361 sediment in the experiments resulted in limited changes to the distribution of
9 362 individuals in the presence of flow. The large, homogenous gravel matrix used in the
10 363 experiments most likely helped maintain open interstitial spaces despite
11 364 sedimentation of the benthic layer (0-50mm). In addition, it is likely that some
12 365 movement of sediment from the surface into lower sections of the experimental
13 366 column occurred during the experiments due to the effect of gravity and the activity
14 367 of individuals. This effectively maintained the interstitial spaces and porosity (Xu et
15 368 al., 2012) thus allowing faunal movement when only one layer of the column was
16 369 treated. Substratum composition and particle size have been widely acknowledged
17 370 as playing a pivotal role in the influence of fine sediment on invertebrate
18 371 communities, with heterogeneous river beds cited as having the greatest clogging
19 372 potential (Weigelhofer & Waringer, 2003). Consequently, the coarse grained
20 373 sedimentary characteristics of the substrates which are subject to sedimentation may
21 374 determine the effects experienced by the ecosystem.

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37 375
38 376 Reductions of interstitial pore space have been reported to limit the ability of fauna to
39 377 migrate within the hyporheic zone, most notably larger bodied invertebrates
40 378 (Williams and Hynes, 1974; Gayraud & Phillippe, 2001). However, in this study the
41 379 effect of body size was not considered as the varying pattern of hydrological
42 380 exchange would have confounded the results (with the maximum number of
43 381 individuals being recorded at opposing ends of the columns for upwelling and
44 382 downwelling conditions in most of the experimental runs). Further experimental
45 383 studies focussed on specific patterns of hydrological exchange would be required to
46 384 enable the effect of body size or other morphological traits on the ability of fauna to
47 385 utilize interstitial spaces to be examined (see Descloux et al., 2014).

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56 387 The results from this study suggest that sedimentation / colmation of the hyporheic
57 388 zone has the potential to effectively disconnect it from benthic sediments and
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3 389 macrofaunal use. This disconnecting off effect may prevent the hyporheic zone
4 390 acting as a refugium during adverse conditions in the surface stream (Wood et al.,
5 391 2010) potentially limiting stream productivity and reducing ecosystem resilience
6 392 (Boulton, 2007; 2010). These results provide insights that support *in-situ*
7 393 observational studies regarding the deleterious effects of sedimentation on
8 394 macroinvertebrates, with reductions in the abundance and diversity of invertebrates
9 395 with increasing depth recorded (Richards & Bacon, 1994; Bo et al., 2007; Descloux
10 396 et al., 2013).

11 397
12 398 The approach applied in this study represents a novel experimental design which
13 399 can be easily replicated and adapted to enable the the effects of sedimentation and /
14 400 or patterns of hydrological exchange on other specific taxa or combinations of taxa
15 401 forming the macroinvertebrate community to be examined. However, care is required
16 402 when applying the results to other aquatic invertebrate taxa and the wider community.
17 403 This study examined a single taxon, however the findings from a number of *in-situ*
18 404 studies do suggest a similar response to hyporheic sedimentation for other mobile
19 405 taxa. Only tube building Chironomidae and burrowing Oligochaeta have been widely
20 406 reported to thrive on the presence of high volumes of fine sediment within the
21 407 hyporheic zone (Zweig & Rabeni, 2001; Weigelhofer & Waringer, 2003; Sarriquet et
22 408 al., 2007). In addition, the current experiments were undertaken under highly
23 409 controlled conditions. In the natural environment, physical conditions and water
24 410 quality will probably differ significantly between upwelling and downwelling flow
25 411 (Olsen & Townsend, 2003; Krause et al., 2011). This strong physio-chemical
26 412 gradient may exert a strong influence on the distribution of both benthic and
27 413 hyporheic invertebrate communities and thus may influence invertebrate response
28 414 (Pepin & Hauer, 2002; Davy-Bowker et al., 2006). There is clearly a need for
29 415 additional experimental studies to gain a better understanding of the factors
30 416 controlling the use of the hyporheic habitats by benthic fauna and to quantify the
31 417 influence of sedimentation on macroinvertebrate communities.

32 418

33 419 **Acknowledgements**

34 420 KLM gratefully acknowledges the support of a Nuffield Undergraduate Research
35 421 Bursary (URB/39419) to undertake the research entitled 'The influence of fine
36 422 sediment deposition and clogging on macroinvertebrate utilization of benthic and

1
2
3 423 hyporheic sediments'. PJW acknowledges the support of the British Cave Research
4 424 Association Cave Science and Technology Research Initiative grant to develop the
5 425 experimental hypogean laboratory. Thanks to Mark Szegner for help with the
6 426 production of the figures.
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Copy for Review

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617 **List of Figures**

618

619 Figure 1. Fine sediment treatments applied to sections / layers of substratum
620 columns (A – 0-50 mm; B – 50-100 mm; C – 100-150 mm; D – 150-200 mm; and E –
621 200-250 mm) during experiments: 1. Open gravel framework at all layers (control
622 conditions); 2. Benthic sedimentation with the equivalent of 3 kg m²; 3. Benthic
623 sedimentation with the equivalent of 5 kg m²; 4. Hyporheic sedimentation of one
624 layer (100-150 mm) with the equivalent of 3 kg m²; 5. Hyporheic sedimentation of
625 three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and 6.
626 Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m².

627

628 Figure 2. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section
629 of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-
630 250 mm) under no hydrological exchange (no flow) conditions: a. Open gravel
631 framework at all layers (control conditions); b. Benthic sedimentation with the
632 equivalent of 3 kg m²; c. Benthic sedimentation with the equivalent of 5 kg m²; d.
633 Hyporheic sedimentation of one layer (100-150 mm) with the equivalent of 3 kg m²;
634 e. Hyporheic sedimentation of three layers (50-200mm) with the equivalent of 3 kg
635 m² applied to each layer; and f. Benthic and hyporheic sedimentation (all layers) with
636 the equivalent of 3 kg m². Sections within the column where the number of
637 individuals were not significantly different are indicated with the same letter (Fisher's
638 LSD, P <0.05).

639

640 Figure 3. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section
641 of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-
642 250 mm) during downwelling flow conditions: a. Open gravel framework at all layers
643 (control conditions); b. Benthic sedimentation with the equivalent of 3 kg m²; c.
644 Benthic sedimentation with the equivalent of 5 kg m²; d. Hyporheic sedimentation of
645 one layer (100-150 mm) with the equivalent of 3 kg m²; e. Hyporheic sedimentation
646 of three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and
647 f. Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m².
648 Sections within the column where the number of individuals were not significantly
649 different are indicated with the same letter (Fisher's LSD, P<0.05).

650

651 Figure 4. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section
652 of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-
653 250 mm) during upwelling flow: a. Open gravel framework at all layers (control
654 conditions); b. Benthic sedimentation with the equivalent of 3 kg m²; c. Benthic
655 sedimentation with the equivalent of 5 kg m²; d. Hyporheic sedimentation of one
656 layer (100-150 mm) with the equivalent of 3 kg m²; e. Hyporheic sedimentation of
657 three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and f.
658 Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m².
659 Sections within the column where the number of individuals were not significantly
660 different are indicated with the same letter (Fisher's LSD, P<0.05).

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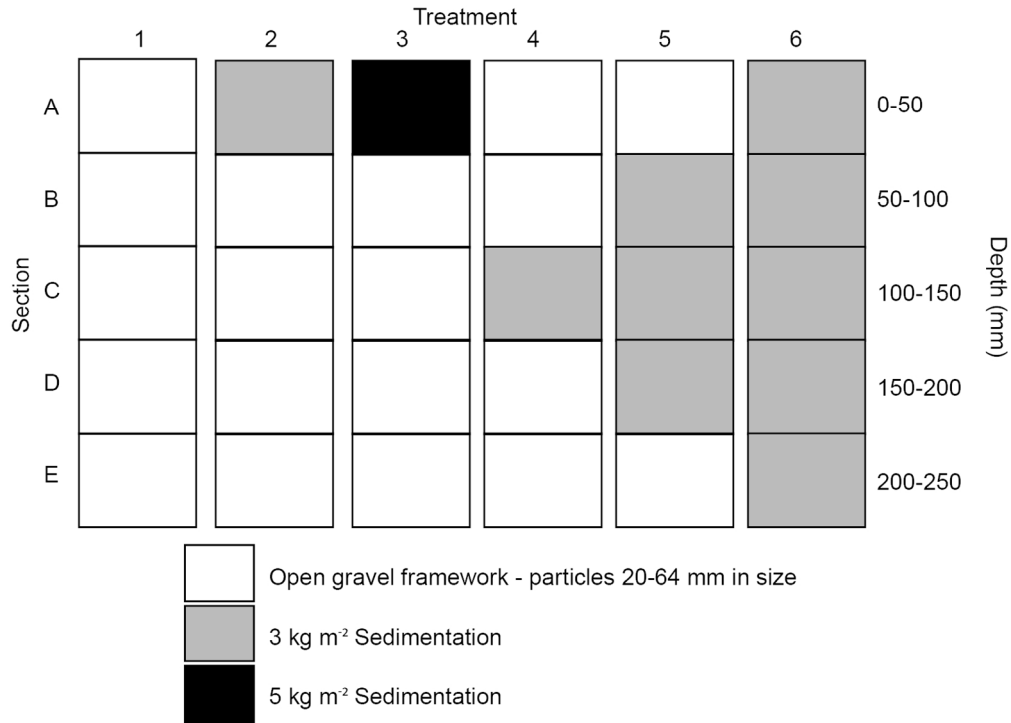


Figure 1. Fine sediment treatments applied to sections / layers of substratum columns (A – 0-50 mm; B – 50-100 mm; C – 100-150 mm; D – 150-200 mm; and E – 200-250 mm) during experiments: 1. Open gravel framework at all layers (control conditions); 2. Benthic sedimentation with the equivalent of 3 kg m²; 3. Benthic sedimentation with the equivalent of 5 kg m²; 4. Hyporheic sedimentation of one layer (100-150 mm) with the equivalent of 3 kg m²; 5. Hyporheic sedimentation of three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and 6. Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m².

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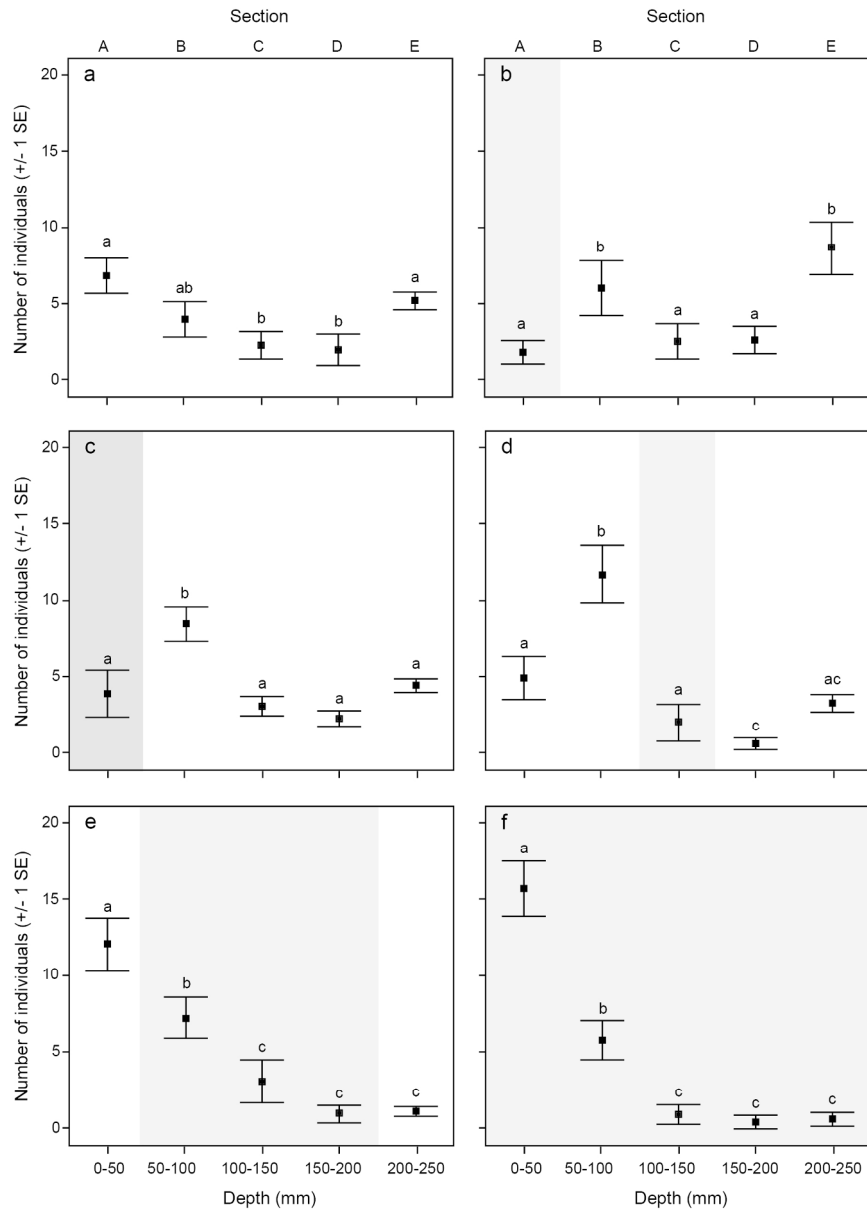


Figure 2. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-250 mm) under no hydrological exchange (no flow) conditions: a. Open gravel framework at all layers (control conditions); b. Benthic sedimentation with the equivalent of 3 kg m²; c. Benthic sedimentation with the equivalent of 5 kg m²; d. Hyporheic sedimentation of one layer (100-150 mm) with the equivalent of 3 kg m²; e. Hyporheic sedimentation of three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and f. Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m². Sections within the column where the number of individuals were not significantly different are indicated with the same letter (Fisher's LSD, P <0.05).

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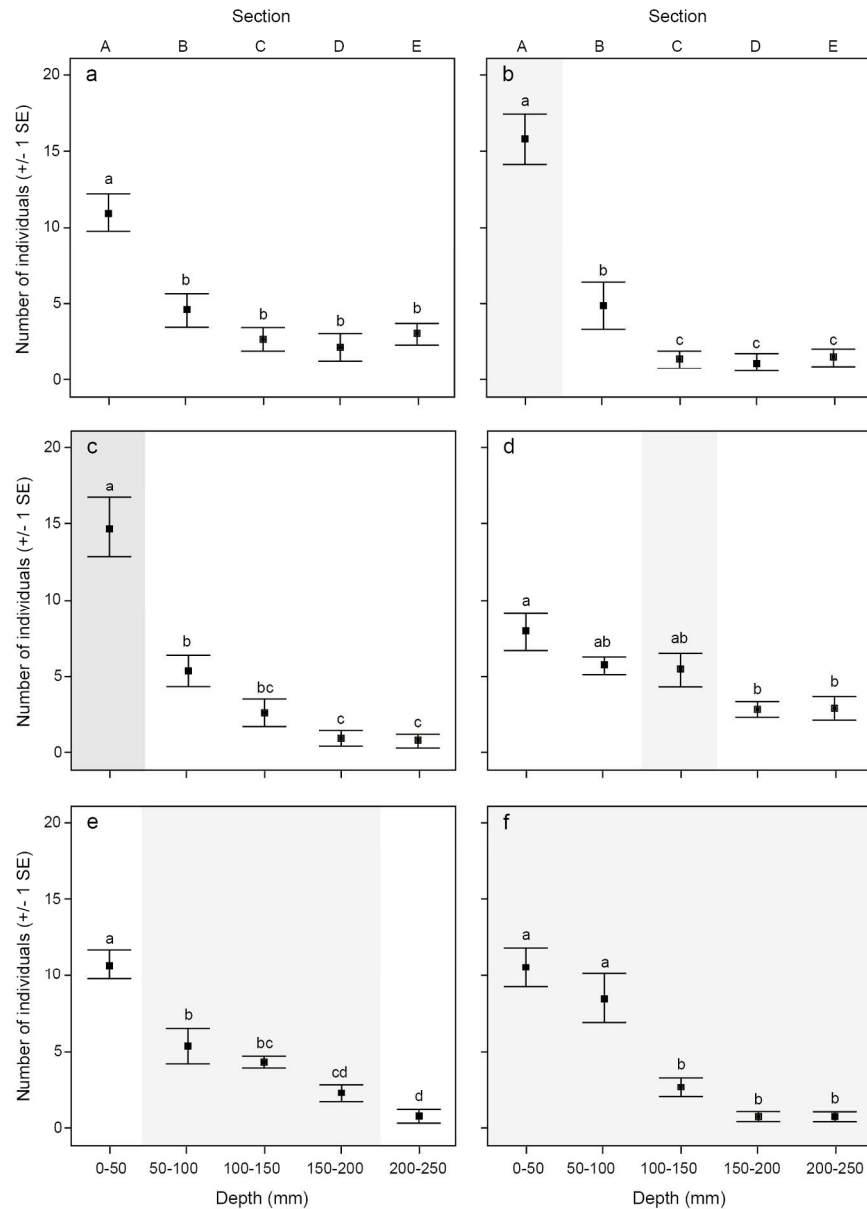


Figure 3. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-250 mm) during downwelling flow conditions: a. Open gravel framework at all layers (control conditions); b. Benthic sedimentation with the equivalent of 3 kg m⁻²; c. Benthic sedimentation with the equivalent of 5 kg m⁻²; d. Hyporheic sedimentation of one layer (100-150 mm) with the equivalent of 3 kg m⁻²; e. Hyporheic sedimentation of three layers (50-200mm) with the equivalent of 3 kg m⁻² applied to each layer; and f. Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m⁻². Sections within the column where the number of individuals were not significantly different are indicated with the same letter (Fisher's LSD, P<0.05).

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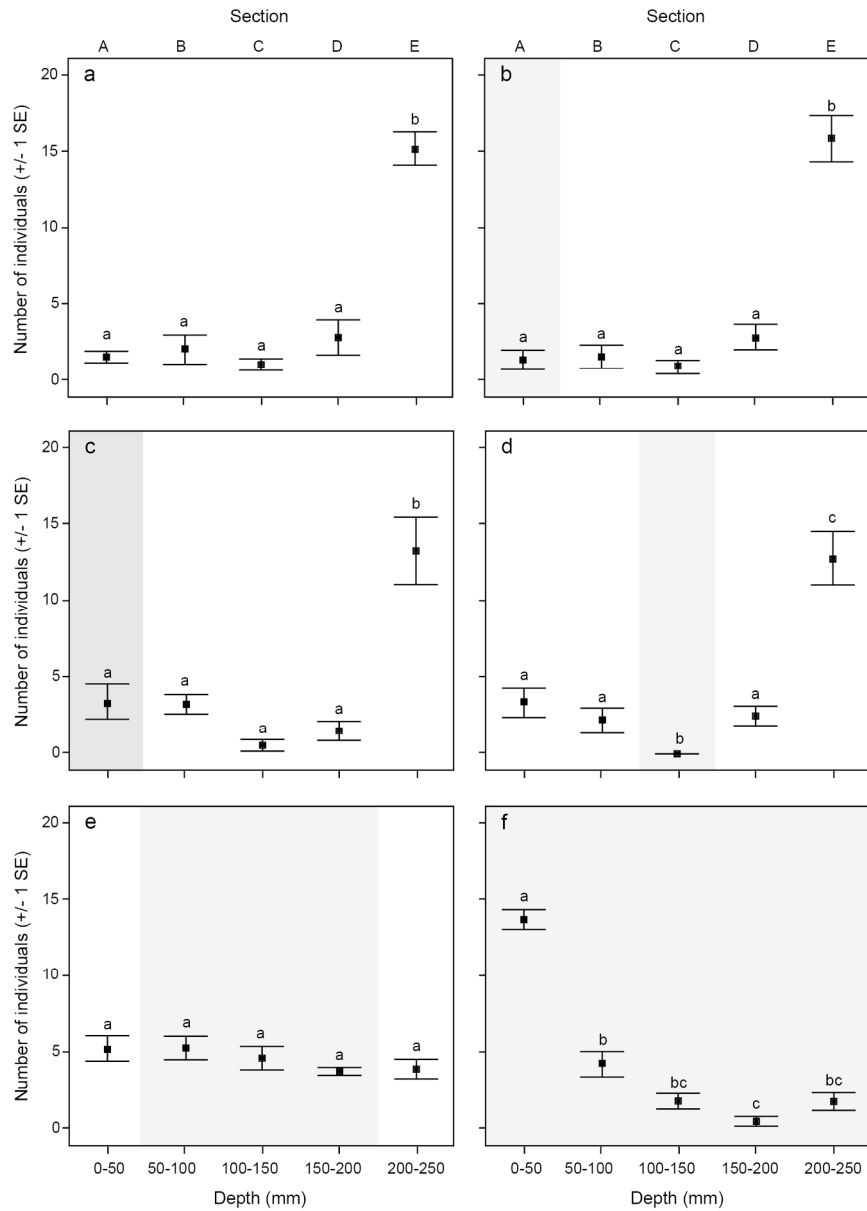


Figure 4. Mean number of *Gammarus pulex* (+/- 1SE) recorded within each section of the sediment column (0-50 mm; 50-100 mm; 100-150 mm; 150-200 mm and 200-250 mm) during upwelling flow: a. Open gravel framework at all layers (control conditions); b. Benthic sedimentation with the equivalent of 3 kg m²; c. Benthic sedimentation with the equivalent of 5 kg m²; d. Hyporheic sedimentation of one layer (100-150 mm) with the equivalent of 3 kg m²; e. Hyporheic sedimentation of three layers (50-200mm) with the equivalent of 3 kg m² applied to each layer; and f. Benthic and hyporheic sedimentation (all layers) with the equivalent of 3 kg m². Sections within the column where the number of individuals were not significantly different are indicated with the same letter (Fisher's LSD, P<0.05).

154x216mm (300 x 300 DPI)