

1 RRH: HYDESKOV ET AL. – LEAD (Pb) EXPOSURE AND ITS HEALTH EFFECTS IN
2 WILD MAMMALS

3 **A Global Systematic Review of Lead (Pb) Exposure and its Health Effects in Wild**

4 **Mammals**

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6 **Helle B. Hydeskov,^{1,5} Jon M. Arnemo,^{2,3} Chris Lloyd Mills,⁴ Louise K. Gentle,¹ and**

7 **Antonio Uzal¹**

8

9 ¹ School of Animal, Rural and Environmental Sciences, Nottingham Trent University,

10 Brackenhurst Campus, Brackenhurst Lane, Southwell, NG25 0QF, UK

11 ² Department of Forestry and Wildlife Management, Faculty of Applied Ecology, Agricultural

12 Sciences and Biotechnology, Inland Norway University of Applied Sciences, Campus Evenstad,

13 Anne Evenstads Vei 80, 2480 Koppang, Norway

14 ³ Department of Wildlife, Fish, and Environmental Studies, Faculty of Forest Sciences, Swedish

15 University of Agricultural Sciences, Skogsmarksgränd, 90736 Umeå, Sweden

16 ⁴ School of Science and Technology, Nottingham Trent University, Clifton Campus, Clifton

17 Lane, Nottingham, NG11 8NS, UK

18

19 ⁵ Corresponding author (email: hbhvet@gmail.com)

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21 ABSTRACT: Lead (Pb) is a toxic non-essential metal, known mainly for causing poisoning of
22 humans and wild birds. However, little is known about Pb exposure and its associated health
23 effects in wild mammals. We conducted a global systematic literature review to identify peer-
24 reviewed studies published on Pb exposure in wild mammalian species and the health effects
25 they identified. In total, 183 studies, conducted in 35 countries and published over 62 yr (1961-
26 2022), were included in the review. Only 6% (11/183) of the studies were conducted in
27 developing countries. Although 153 mammalian species were studied, most studies focused on
28 species that are easy to access (i.e., hunted species and small mammals that are easy to trap).
29 Therefore, carnivores and scavengers were less frequently studied than herbivores and
30 omnivores. Despite all studies reporting Pb concentrations, only 45 (25%) studies investigated
31 health effects and, of these 45 studies, only 28 (62%) found any health effect in 57 species. All
32 health effects were negative, and ranged from subclinical effects to fatality. Methodologies of Pb
33 sampling and quantification, and reporting of results, varied widely across the studies, making
34 both Pb concentrations and health effects difficult to compare and evaluate. Thus, there is a need
35 for more research on Pb exposure and its health effects on wild mammals, especially as
36 carnivores and scavengers could be used as sentinels for ecosystem health.

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38 *Key words:* Bioaccumulation, Mammalia, One Health, sentinel species, toxicology, wildlife
39 population health

INTRODUCTION

40

41 Lead (Pb) is a toxic, non-essential metal that has no known biological function for living
42 organisms, contaminates the environment, and causes negative health effects in humans and
43 other animals (Ma 2011; European Chemicals Agency 2018; Lanphear et al. 2018; Pain et al.
44 2019). It occurs naturally in the Earth's crust, and has historically been used in many products,
45 including water pipes, paints, gasoline, ammunition, aviation fuel, motor vehicle batteries,
46 glassware and cosmetics (Bergdahl and Skerfving 2022). These anthropogenic activities have
47 increased the amount of environmental Pb to around 1,000 times the natural levels (Renberg et
48 al. 2001), thereby amplifying the potential environmental exposure of wildlife to Pb. Despite
49 legislation banning or reducing the use of Pb, it is still present in many products. For example,
50 very few countries, states or regions have banned all Pb-based ammunition (Sonne et al. 2022)
51 and, according to the World Health Organization (WHO) as of June 2022, only 45% of countries
52 had made legally-binding controls on the production, import, sale, and use, of Pb-based paints
53 (WHO 2022a).

54 Children are generally more at risk of health effects from Pb exposure than adults, as
55 children absorb more ingested Pb from their gastrointestinal system (WHO 2022b). Multiple
56 organ systems, including the nervous, renal, cardiovascular, hematological, immune and
57 reproductive systems, are negatively affected by Pb. The health effects of Pb exposure range
58 from chronic and subclinical, to acute and fatal (Agency for Toxic Substances and Disease
59 Registry 2020; Bergdahl and Skerfving 2022). Furthermore, there is no known safe level of Pb
60 exposure in humans (WHO 2022b). Despite this, concentrations of concern vary between health
61 authorities. For example, the Centers for Disease Control and Prevention (CDC) in the USA
62 currently have a "blood lead reference value (BLRV)" of 35 µg/L, which is used to identify

63 children with high blood Pb concentrations (the 2.5% with the highest concentration) that require
64 additional actions to be taken by health care professionals (CDC 2022). The European Food
65 Safety Authority (EFSA) uses a different system and has set lower confidence limits of 12, 15,
66 and 36 µg/L as benchmark doses for developmental neurotoxicity, effects on the prevalence of
67 chronic kidney disease, and effects on systolic blood pressure, respectively (EFSA Panel on
68 Contaminants in the Food Chain (CONTAM) 2013). Furthermore, concentrations of concern set
69 by different authorities may vary over time: Between 2012 and 2021, CDC's BLRV was 50
70 µg/L; prior to 2012, it was 100 µg/L and called the "level of concern in children" (CDC 2022).
71 With this continuing lowering of the CDC's BLRV, it is clear that both health authorities agree
72 that even very small concentrations of Pb can negatively affect human health, in alignment with
73 the WHO's understanding of Pb exposure (WHO 2022b).

74 Research on Pb exposure and health effects in wildlife has often been undertaken on
75 avian species, where the first poisonings were recognized nearly 150 yr ago (Calvert 1876). In
76 birds, Pb exposure often leads to multisystemic clinical disease or even death, especially in
77 waterfowl, scavenging birds and predatory birds (e.g., Pain et al. 2019). Consequently, Pb
78 exposure has caused population declines, for example in the scavenging California condor
79 (*Gymnogyps californianus*), a scavenging species (Green et al. 2008). Although toxicity
80 thresholds in different avian species have previously been proposed (Franson and Pain 2011),
81 scientists have recently suggested stopping this practice as it gives the impression that some
82 levels of Pb are acceptable (Pain et al. 2019).

83 Previous reviews on Pb in mammals have focused primarily on animals in laboratory
84 settings or on domestic species. For example, Ma (2011) showed that humans and laboratory
85 mammals have similar susceptibility to Pb exposure (i.e., the same Pb dose-health effects

86 relationship). In domestic mammals, young cattle appear most prone to Pb poisoning, probably
87 due to their tendency to lick different items (Constable et al. 2017). However, young cattle may
88 simply absorb Pb more efficiently than adults, as seen in humans (WHO 2022b).

89 There is a lack of a global systematic literature review focusing on Pb exposure and
90 health effects in wild mammalian species. Synthesizing this scientific evidence is important to
91 understand the breadth of wild mammalian species affected by Pb across the world, as there is no
92 reason to expect wild mammals to be less at risk from Pb exposure than any other vertebrates,
93 including humans.

94 Our article was designed to provide a global systematic review of the literature on
95 environmental Pb exposure and health effects in wild mammals, including assessment of gaps
96 and biases in the research, and suggestions on how to advance knowledge of the effects of Pb,
97 particularly in identifying the types of species that may best indicate environmental Pb exposure.
98 The outcomes of this review are important because they can be used to inform future studies,
99 influence policymakers, and contribute towards relevant environmental legislation.

100 MATERIALS AND METHODS

101 Systematically identifying peer-review studies to include

102 This systematic review followed the workflow proposed by the Preferred Reporting Items
103 for Systematic reviews and Meta-Analyses (PRISMA) 2020 statement (Page et al. 2021).

104 A literature search for peer-reviewed scientific articles on Pb exposure in wild mammals was
105 conducted using the database Scopus on 21 June 2022. Scopus was chosen over other well-
106 known databases and online search engine, such as Web of Science, PubMed, or Google Scholar
107 due to the ability of Scopus to allow the chemical lead (Pb) as a search term, an option not
108 provided by other databases or online search engines. In addition, Scopus permitted the

109 exclusion of articles containing the word “lead” when used as a verb in the absence of any
110 occurrence of lead as a chemical in the source. Furthermore, Scopus only indexed peer-reviewed
111 articles and did not have restrictions on the length of search queries.

112 In Scopus, different query combinations of the following searches were used:

113 CASREGNUMBER (7439-92-1)/CHEMNAME (lead); wild/wildlife/free-living/"free
114 living"/free-ranging/"free ranging"/semi-wild/"semi wild"/free-roaming/"free
115 roaming"/street/feral; captive/zoo/domestic/domesticated/semi-captive/"semi captive"/semi-
116 domestic/"semi domestic"/semi-domesticated/"semi domesticated";
117 exposure/expose/exposed/exposing/poisoning/poison/poisoned; keywords for mammals, and
118 scientific nomenclature of all mammalian orders, families and genera taken from the
119 International Union for the Conservation of Nature (IUCN) Red List of Threatened Species
120 (IUCN 2021); details are given in the Supplementary Materials.

121 The first author conducted the assessment of all records and screening of studies. After
122 identifying records via Scopus, records were checked for duplicates. Unique records were then
123 screened by reading the title and abstract. The selected studies were thoroughly screened by
124 reading the entire article. Articles on Pb exposure in wild mammals that were cited in the
125 reference sections of the Scopus articles passing the full screening were also screened to identify
126 additional publications not found by the original literature searches (i.e., snowball search).
127 Inclusion criteria included the requirement that studies must be peer-reviewed original research
128 studies reporting Pb concentrations in tissues or body fluids in a non-domestic mammal (called
129 “wild mammals” in this review) species. Studies written in any language were included and
130 language was not used as an exclusion criterion. Studies reporting Pb concentrations in captive
131 wild mammals outside laboratory settings were included to illustrate the variety of wild

132 mammalian species documented to be susceptible to Pb exposure from their environment.
133 Exclusion criteria included articles not reporting Pb concentrations in wild mammalian tissues or
134 blood, studies using purely experimental settings, and studies reporting Pb concentrations in free-
135 ranging domestic mammalian species. Data from studies were excluded if the study reported Pb
136 concentration in tissues or body fluids knowingly contaminated with Pb-based ammunition (i.e.,
137 from killing the animal). Some studies reported non-standard units (i.e., liquid concentrations for
138 tissues or units unknown to the authors of this study). In these cases, or where there was a
139 mismatch between data reported in text and tables, the corresponding author was contacted via
140 the email provided in the publication and given a maximum of 30 d to respond. Without a
141 response, either the data in question or the entire study was excluded.

142 **Data extracted from the studies**

143 For every study included, the following information was extracted where available:
144 Taxonomy (order, family, common species name, scientific species name); continent and country
145 where the study was conducted; description of known pollution sources nearby the study site;
146 type of living condition (free-ranging, captive or semi-captive); year of sample collection;
147 season; age; sex; tissue; sample size, Pb analysis method; reporting in wet weight or dry weight
148 concentrations (where both were reported, dry weight concentration was preferred); and Pb
149 concentration arithmetic mean (geometric means and arithmetic weighted means were excluded),
150 standard deviation, standard error, median, minimum value and maximum value. The Pb
151 concentrations were converted to $\mu\text{g}/\text{kg}$ or $\mu\text{g}/\text{L}$ where applicable, for comparison. Whether the
152 study had investigated and identified any health effects of Pb was also recorded. To investigate
153 in which part of the mammalian food web Pb exposure is most frequently reported, the feeding
154 type of the mammalian species (i.e., herbivore, omnivore or carnivore), and whether or not the

155 species scavenges on animal material, was determined by following species descriptions
156 available in a recently published mammalian textbook (Wilson and Mittermeier 2009-2019).
157 Articles from Russia were allocated to the European or Asia continent based on the geographical
158 location of the sample collection. To investigate potential economic resource biases in the
159 publications, the Organisation for Economic Co-operation and Development's designation of
160 developed and developing countries was used (Organisation for Economic Co-operation and
161 Development 2006).

162 Different parts of a tissue or body fluid (e.g., whole kidney versus kidney cortex or whole
163 blood versus plasma) were considered separate tissues for the purpose of the review.
164 Additionally, cortical bone and antler were differentiated, as cortical bone is permanent bone,
165 whereas antler is deciduous bone (representing only short-term or annual exposure).

166 **RESULTS**

167 **Characteristics of the studies**

168 A total of 254 records were retrieved from Scopus (Fig. 1), with no duplicates identified.
169 After the initial screening of the title and abstract, 125 studies remained for full screening,
170 resulting in 69 studies from the Scopus search queries meeting the inclusion criteria.
171 Additionally, 114 studies were identified for inclusion through a snowball search, giving a total
172 of 183 studies for systematic review (see Supplementary Materials).

173 The studies were published over 62 yr (1961 to June 2022), with most of the studies
174 (158/183, 86%) published in the second half of this period (1992 to 2022) (Fig. 2). The majority
175 of studies (162/183, 89%) were conducted in Europe (108 studies, 59%) and North America (54
176 studies, 30%); some studies were conducted in multiple countries (but always within the same
177 continent). The US was the country where most studies were conducted (38 studies), followed by

178 Canada (17 studies), Poland (16 studies), Germany (12 studies) and the UK (11 studies). Only
179 6% (11/183) of the studies were conducted in developing countries (Fig. 3).

180 **Lead (Pb) exposure**

181 Lead was present in all mammalian species studied, with concentrations reported in a
182 total of 153 wild mammalian species belonging to 45 families and 11 orders. The most studied
183 order, family and species, were *Cetartiodactyla*, *Cervidae* and red deer (*Cervus elaphus*), with
184 67, 46 and 20 studies, respectively (Table 1). In terms of feeding type, herbivores, omnivores and
185 carnivores were researched in 92, 79 and 49 studies, respectively; 49 studies researched
186 scavenging species.

187 A total of 33 different tissues and body fluids were tested. The five most studied tissues
188 were the liver (123 studies), kidney (92 studies), skeletal muscle (54 studies), bone (32 studies)
189 and blood (25 studies). At 1,506,000 µg/kg dry weight, the highest Pb concentration reported
190 across all 11 mammalian orders studied was recorded in the kidney of a Northern short-tailed
191 shrew (*Blarina brevicauda*; Table 2).

192 **Health effects of lead (Pb)**

193 A total of 45/183 (25%) studies investigated the health effects of Pb. Of these, 28 (62%)
194 studies identified a negative health effect, in 57 species. No studies identified a positive health
195 effect. Negative health effects were reported in 32 herbivores, 21 omnivores and four carnivores
196 and ranged in severity from ‘interference with developing permanent teeth’ to ‘fatality’
197 (Table 3).

198 **Variation in methodology**

199 The 183 studies used a large variety of sampling approaches and Pb analysis techniques,
200 and reported the results in different ways, making it hard to compare findings. For example,

201 some studies only reported mean values, while others reported mean, standard deviation or
202 standard error or both, median and range of Pb concentration. Full details on sex and age of the
203 animals tested, and on whether the sampling was undertaken in an area with known Pb pollution,
204 were not provided in 158 (86%), 134 (73%), and 101 (55%) studies, respectively.

205 **DISCUSSION**

206 **Coverage of lead (Pb) exposure in wild mammals in the literature**

207 The number of publications of Pb studies on wild mammalian species has increased over
208 the last three decades. Nevertheless, the frequency of Pb publications on wild avian species still
209 appears to be much higher, indicated by the large number of avian reviews published in recent
210 years (e.g., Pain et al., 2019; Monclús et al., 2020; Ives et al., 2022).

211 Pb concentrations were reported in 153 wild mammalian species, highlighting that this
212 class of animals is exposed to Pb on a large scale. Despite this clear evidence of Pb exposure
213 across many taxa, based on IUCN taxonomic classification, only 41% (11/27) of orders, 28%
214 (45/162) of families and 3% (153/5,968) of species were studied, which evidences a large group
215 of mammals with no information on Pb exposure (IUCN 2021). Understandably, there has been a
216 tendency to study species that are easiest to acquire (i.e., hunted species and small species that
217 are easy to trap without chemical immobilization). Consequently, the studies identified in this
218 review are skewed towards herbivores and omnivores. It is not possible to determine whether
219 tissues were selected due to easy access, or because the researcher hypothesized that the selected
220 tissue would provide the most information in their study species. However, liver, kidney, skeletal
221 muscle, bone and blood are commonly collected and stored for further investigations during
222 wildlife necropsies (McAloose et al. 2018).

223 **Health effects of lead (Pb) in wild mammals**

224 Only 28 studies, 62% of those in which health effects were investigated, identified
225 negative health effects of Pb in wild mammals. As expected, no positive effects were found.
226 From the breadth of health effects identified and organ systems affected, it is clear that Pb affects
227 many parts of the body in wild mammals, just like in other species (Pain et al. 2019; Bergdahl
228 and Skerfving 2022). Many of the identified health effects would most likely be subclinical, so it
229 is not possible to determine how exactly these impact wild mammalian populations.

230 It is expected that, due to their feeding behaviors, some mammal species are more likely
231 than some bird species to be exposed to smaller amounts of Pb, relative to body mass, but over
232 extended periods of time. For example, waterfowl are known to pick up large numbers of Pb-
233 based shotgun pellets or fishing weights from the bottom of wetland areas when foraging (Pain et
234 al. 2019), yet most wild mammals would not be exposed to such large amounts of Pb at once, as
235 they forage using different methods (e.g., *Carnivora* species). This difference in feeding
236 behavior would reduce the risk of acute poisonings, seen in many waterfowl, scavenging and
237 predatory bird species, but may cause more chronic and often subclinical health effects (e.g., on
238 the nervous and immune systems). Chronic effects tend to be much less apparent than acute
239 effects in free-ranging wildlife, so may be overlooked unless investigated explicitly. Ecke et al.
240 (2017) concluded that sublethal Pb exposure in golden eagles (*Aquila chrysaetos*) might impair
241 flight performance and thus increase the risk of mortality from other causes. Yet, other eagle
242 studies could not find evidence that subclinical Pb concentrations are linked to “increased risk of
243 trauma” (Isomursu et al. 2018) or “other causes of mortality” (Franson and Russell 2014). These
244 examples demonstrate the difficulty in investigating subclinical and chronic health effects in
245 wildlife, especially when studies do not classify effects similarly. Furthermore, ethical permits to

246 conduct Pb toxicological effect studies on wild mammals and other wildlife are, rightly, unlikely
247 to be granted. Therefore, subclinical effects are difficult to confirm in wildlife.

248 Although the impacts of Pb exposure on human, laboratory mammalian, and wild avian
249 health are well-documented, there is still much to learn about the impacts on wild mammal
250 health. In contrast to important infectious diseases in animals, the World Organisation for
251 Animal Health does not have any reference laboratories for any non-infectious diseases such as
252 Pb poisoning, which could guide or help standardize how the diagnostic steps should be
253 conducted (World Organisation for Animal Health 2022). Potential strategies to research the
254 health effects of Pb on wild mammals include behavioral studies, reproductive success analyses,
255 and immune system function assessments; these are all challenging to conduct in free-ranging
256 animals. Research which would be easier to achieve in free-ranging settings includes correlations
257 between blood Pb concentrations and hematological and biochemical parameters; investigations
258 of blood and tissue Pb concentrations to understand the distribution of Pb in the body; and
259 investigations of tissue pathologies in Pb-exposed populations. Ideally, studies should use the
260 most sensitive techniques available for measuring Pb, such as inductively-coupled plasma mass
261 spectrometry (Apostoli 2002) and report the results in dry weight for easier comparisons.
262 Regardless of the approach taken by future studies to investigate Pb exposure and health effects
263 in wild mammals, studies should report details of the study area, sex and age of animals, and
264 analytical methods, as well as providing an appropriate range of descriptive statistics (e.g.,
265 sample sizes, standard deviation or standard error), which would facilitate future evidence-based
266 policies.

267 **Methodologies, kinetics and thresholds**

268 Our systematic review found a lack of consistency in estimating Pb concentrations, health
269 effects and how this is reported. The different Pb analytical methods used in the literature
270 reviewed (e.g., atomic absorption spectroscopy and inductively coupled plasma mass
271 spectrometry) vary in sensitivity and differ in susceptibility to matrix effects (Apostoli 2002).
272 Therefore, some studies might have failed to detect Pb if present in low concentrations. In
273 addition, some studies reported tissue concentrations by dry weight and others by wet weight,
274 impeding direct comparison of such concentrations. Furthermore, Pb uptake and storage
275 distribution potentially may differ between species: As there are no studies available on Pb
276 kinetics in wild mammals, this remains unknown. Variations in methodologies and target species
277 make it difficult to compare Pb concentrations between studies and between species.
278 Consequently, it is not possible to establish toxicity thresholds for wild mammals based on the
279 current body of literature. However, although thresholds may be a valuable tool for some
280 professions (e.g., veterinarians in clinical practice), a pragmatic reason for not attempting to
281 establish threshold values for different wild mammalian tissues or species is the risk of creating
282 the false impression of acceptable or safe levels of Pb exposure in wild mammals. Therefore, we
283 recommend ceasing the use of threshold values for Pb concentrations in wildlife research, as
284 already suggested by Pain et al. (2019).

285 **Designing future studies on lead (Pb) exposure and health effects**

286 From a One Health perspective, human, animal and ecosystem health are interlinked.
287 Whilst it would be beneficial to study Pb concentrations in species at differing trophic levels in
288 the wild, the cost of periodic screening programs means that these are likely to be restricted to
289 only a few selected species. Since Pb bioaccumulates in both plants and animals (Clemens 2006;
290 Radomyski et al. 2018), testing for Pb in carnivorous and scavenging mammalian species might

291 not only aid monitoring of environmental Pb exposure, but also indicate the potential exposure of
292 humans to Pb, since species at these higher trophic levels often have long lifespans like humans
293 (O'Brien et al. 1993). Using higher trophic-level species as sentinels is already being used in
294 avian Pb research (e.g., Monclús et al. 2020), and in mammalian research on infectious diseases
295 and antimicrobial resistance (e.g., Millán et al. 2014; Sacristán et al. 2020). Another benefit of
296 studying top predatory and/or scavenging mammals, which are often iconic and charismatic
297 (Albert et al. 2018), is that these mammals might attract more attention from stakeholders and
298 policymakers involved in the regulation of Pb, a phenomenon already seen in conservation, with
299 charismatic species receiving most attention (Sitas et al. 2009).

300 To assess or monitor short-term and recent Pb exposure using reduced sampling effort
301 and non-invasive sampling, antlers or hairs might be suitable tissues to consider in the future.
302 Antlers represent around 6 mo's worth of Pb exposure, spring to autumn, and have previously
303 been used to assess Pb contamination (Ludolphy et al. 2022). Antlers do, however, introduce
304 potential biases into a Pb monitoring program. For most cervid species, antlers are only available
305 from adult and subadult males, thus detecting differences between age classes and sexes would
306 not be possible using this tissue. Hair is often easily available even from museum specimens
307 (i.e., on pelts), and it can be collected from free-ranging mammals non-invasively using hair
308 traps. The period of Pb exposure can be estimated using laser ablation techniques provided that
309 potential Pb surface contamination is removed, the particular species' hair growth cycle is
310 known, and the sampling is carried out in a consistent manner (Sela et al. 2007).

311 To gain information on where Pb is stored in a particular wild mammalian species,
312 researchers would have to analyze multiple tissues and blood all obtained from the same
313 individual at the same time (Krone 2018; Bergdahl and Skerfving 2022). However, obtaining this

314 quantity of samples would likely only be possible in freshly dead or mildly autolyzed individuals
315 (McAloose et al. 2018). Additionally, extra expenses from analyzing multiple samples from each
316 individual could be cost-prohibited. Therefore, more controlled studies on the storage and
317 clinical pathology of Pb in captive individuals could help to understand the health effects of Pb
318 in wild mammals.

319 With studies written in any language being included, this review highlights that few
320 studies on Pb exposure in wild mammals have been conducted in developing countries. It is
321 important to note that detection of studies in general always depends on how journals have
322 indexed their articles in various databases, therefore it cannot be ruled out that studies written in
323 other languages than English will have been missed. Moreover, it is important to consider that
324 fewer studies published in developing countries does not mean the problem is less prevalent in
325 these countries. Indeed, the lack of studies could be explained by lower access to research
326 funding (Kpokiri et al. 2022). On the contrary, environmental exposure to Pb might be greater in
327 developing countries due to fewer legislative controls on Pb use in products, as seen with Pb-
328 based paints (WHO 2022a). Indeed, children in developing countries are more exposed to Pb
329 than children in developed countries (Hwang et al. 2019; Bergdahl and Skerfving 2022).
330 Therefore, more research on Pb exposure in wild mammals must be conducted and published in
331 these countries to understand the levels of environmental exposure to Pb, the risk to human and
332 animal health, and to study whether legislation on Pb in products affects the Pb present in the
333 environment.

334 As all organisms are exposed to Pb, studies helping to determine the impacts of Pb on
335 different species will add to current knowledge of this toxic metal. Collectively, this increase in
336 knowledge will create stronger evidence to influence policymakers involved in the regulation of

337 Pb in products which continue to act as sources of environmental contamination and be a major
338 One Health issue. Overall, Pb exposure in wild mammalian species must be recognized as an
339 issue, as this systematic review demonstrates. Control and regulation of Pb need to take a One
340 Health approach, not focus on single taxa groups when developing policies.

341 **ACKNOWLEDGMENTS**

342 We thank Victoria Boskett (Research Librarian at Nottingham Trent University) for her
343 help with journal indexing and database search engines.

344 **SUPPLEMENTARY MATERIAL**

345 Supplementary material for this article is online at <http://dx.doi.org/10.7589/JWD-D-23-00055>.

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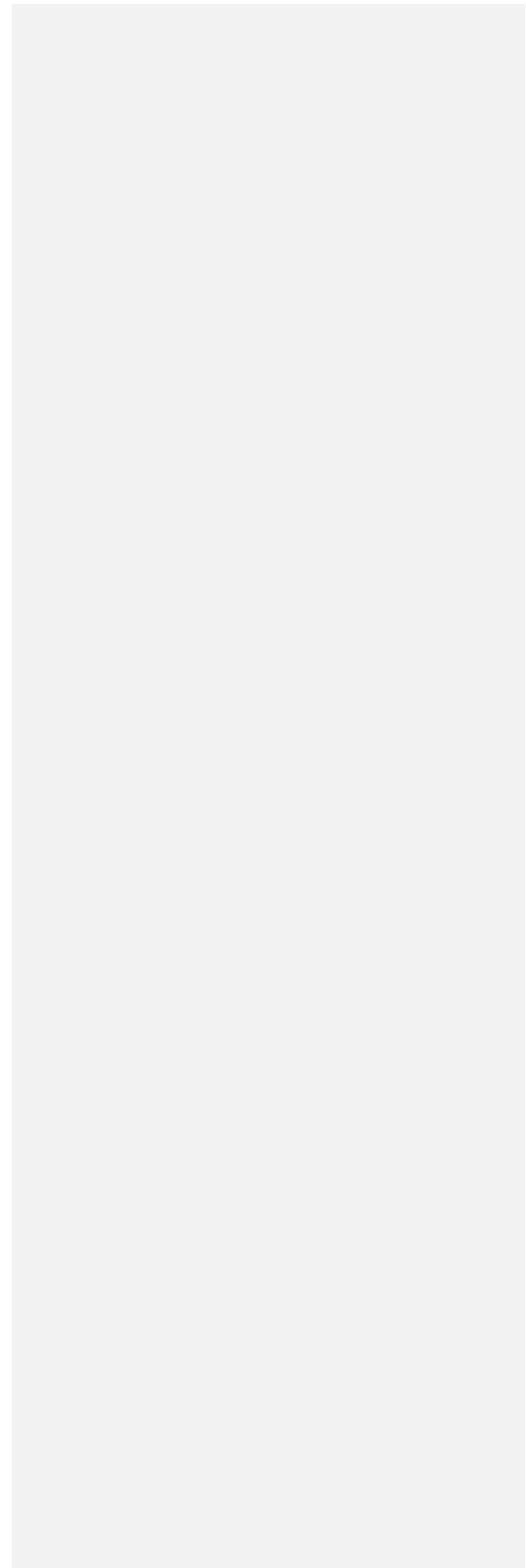


Table 1. Top 5 most studied mammalian orders, families and species by the number of studies published on lead (Pb) exposure in wild mammals between 1961 and June 2022.

Order	No. of studies
Cetartiodactyla	67
Carnivora	55
Rodentia	42
Lagomorpha	15
Eulipotyphla	14
Family	No. of studies
Cervidae	46
Muridae	23
Ursidae	19
Suidae	17
Cricetidae	16
Species	No. of studies
Red deer (<i>Cervus elaphus</i>)	20
European roe deer (<i>Capreolus capreolus</i>)	19
Wild boar (<i>Sus scrofa</i>)	17
Long-tailed field mouse (<i>Apodemus sylvaticus</i>)	12
Brown bear (<i>Ursus arctos</i>)	11

Table 2. Highest individual (where $n=1$) or mean lead (Pb) concentration ($\mu\text{g}/\text{kg}$ or $\mu\text{g}/\text{L}$) reported in each wild mammalian order studied in studies published between 1961 and June 2022. Solid tissue concentrations are reported in either wet weight (w.w.) or dry weight (d.w.).

Order	Concentration	Species	Sample size	Tissue	Location ^a	Country ^b	Reference
<i>Carnivora</i>	861,900 $\mu\text{g}/\text{kg}$ w.w.	Ringed seal (<i>Pusa hispida</i>)	1	Brain	F-R	Canada	(Chan et al. 1995)
<i>Cetartiodactyla</i>	97,300 $\mu\text{g}/\text{kg}$ w.w.	Beluga whale (<i>Delphinapterus leucas</i>)	1	Blubber	F-R	Canada	(Chan et al. 1995)
<i>Chiroptera</i>	500,000 $\mu\text{g}/\text{kg}$ w.w.	Grey-headed flying fox (<i>Pteropus poliocephalus</i>)	2	Liver	Captive	USA	(Zook et al. 1972)
<i>Dasyuromorphia</i>	62 $\mu\text{g}/\text{L}$	Tasmanian devil (<i>Sarcophilus harrisii</i>)	26	Blood	Captive	Australia	(Hivert et al. 2018)
<i>Didelphimorphia</i>	1,300 $\mu\text{g}/\text{kg}$ w.w.	Virginia opossum (<i>Didelphis virginiana</i>)	4	Liver	F-R	USA	(Lewis et al. 2001)
<i>Diprotodontia</i>	2,950 $\mu\text{g}/\text{kg}$ ^c	Common wombat (<i>Vombatus ursinus</i>)	4	Hair	F-R	Australia	(Penrose et al. 2022)
<i>Eulipotyphla</i>	1,506,000 $\mu\text{g}/\text{kg}$ d.w.	Northern short-tailed shrew (<i>Blarina brevicauda</i>)	1	Kidney	F-R	USA	(Stansley and Roscoe 1996)
<i>Lagomorpha</i>	9,000 $\mu\text{g}/\text{kg}$ w.w.	European hare (<i>Lepus europaeus</i>)	9	Muscle	F-R	Austria	(Ertl et al. 2016)
<i>Primates</i>	212,500 $\mu\text{g}/\text{kg}$ w.w.	Stump-tailed macaque (<i>Macaca arctoides</i>)	2	Liver	Captive	USA	(Zook et al. 1972)
<i>Rodentia</i>	672,000 $\mu\text{g}/\text{kg}$ d.w.	Long-tailed field mouse (<i>Apodemus sylvaticus</i>)	5	Bone	F-R	UK	(Johnson et al. 1978)
<i>Sirenia</i>	100 $\mu\text{g}/\text{L}$	American manatee (<i>Trichechus manatus</i>)	4	Blood	Captive	Brazil	(Anzolin et al. 2012)

^a F-R: The wild mammal was from a free-ranging setting. Captive: The wild mammal was from a captive setting.

^b Country in which the study was conducted.

^c No indication of whether the concentration is reported in w.w. or d.w.

Table 3. Negative health effects associated with lead (Pb) exposure in wild mammals by organ system affected in studies published between 1961 and June 2022.

Organ or system affected	Effects found	References
Whole body	Fatality, DNA damage, oxidative stress	Sauer et al. 1970; Shlosberg et al. 1997; Skerratt et al. 1998; da Silva et al. 2000; Brumbaugh et al. 2010; Burco et al. 2012; North et al. 2015; Lazarus et al. 2020

Nervous system	Neurological clinical signs; demyelination of spinal cord and subcortical white matter of brain and optic tracts	Sauer et al. 1970; Zook et al. 1973; Diters and Nielsen 1978; Skerratt et al. 1998
Circulatory system	Decreased ALAD ^a activity; decreased haemoglobin; decreased haematocrit; increased MNEs ^b ; increased total bilirubin; decreased calcium; decreased lactate dehydrogenase	Mouw et al. 1975; Ieradi et al. 1996; Stansley and Roscoe 1996; Rogival et al. 2006; Sánchez-Chardi et al. 2009; Chen et al. 2018
Reproductive system	Testicular structural and functional disruption; abnormal sperm cells; decreased proportion of live sperm	Ieradi et al. 1996; Parelho et al. 2016; Chen et al. 2018
Kidney	Decreased ALAD activity; necrosis and cell degeneration; INIBs ^c in proximal tubular epithelium; oedema; mitochondrial abnormalities; increased weight of kidneys; hyperplasia of tubules, atrophy of glomeruli, interstitial fibrosis; degenerate cells in the lumen of seminiferous tubules; mineralisation and degeneration of proximal tubules; karyocytomegaly; lymphocyte infiltrations; cell debris and change in glomerular aspect and size	Sauer et al. 1970; Zook et al. 1970, 1972; Mouw et al. 1975; Diters and Nielsen 1978; Roberts et al. 1978; Stansley and Roscoe 1996; Skerratt et al. 1998; Ceruti et al. 2002; Damek-Poprawa and Sawicka-Kapusta 2003; Brumbaugh et al. 2010; Tête et al. 2014
Liver	INIBs; necrosis; apoptosis; increased number of pyknotic nuclei; interstitial fibrosis, decreased glycogen content; increased liver-body ratio; steatosis; lymphocyte infiltrations	Sauer et al. 1970; Zook et al. 1970, 1972; Diters and Nielsen 1978; Damek-Poprawa and Sawicka-Kapusta 2003; Sánchez-Chardi et al. 2009; Tête et al. 2014
Teeth	Interference with developing permanent teeth	Rappaport et al. 1975; North et al. 2015

^aALAD: Aminolevulinic acid dehydratase.

^bMNEs: Micronucleated erythrocytes.

^cINIBs: Eosinophilic acid-fast Intranuclear inclusion bodies.

Figure 1. Modified PRISMA 2020 flow diagram for systematic reviews, indicating the number of studies that were included and excluded in the screening process for studies on lead (Pb) exposure in wild mammals.

Figure 2. The number of studies published per year on lead (Pb) exposure in wild mammals included in this systematic review ($n=183$). Studies were published in the period 1961-June 2022.

Figure 3. The number of studies ($n=183$) conducted on lead (Pb) exposure in wild mammals between 1961 and June 2022, per country. Studies were conducted in 35 countries across six continents. Darker color indicates a higher number of studies conducted. Created using Datawrapper (Datawrapper GmbH 2022).